EPA Superfund Record of Decision:

LOWER ECORSE CREEK DUMP EPA ID: MID985574227 OU 01 WYANDOTTE, MI 07/17/1996

Declaration Selected Remedial Alternative for the Lower Ecorse Creek Site Wyandotte, Michigan

Site Name and Location

Lower Ecorse Creek Site North Drive Wyandotte, Michigan 48192

Statement of Basis and Purpose

This decision document presents the selected remedial action for the Lower Ecorse Creek site, in Wyandotte, Michigan, which was chosen in accordance with the requirements of the Comprehensive Environmental Response, Compensation, and Liability Act of 1980 (CERCLA), as amended by the Superfund Amendments and Reauthorization Act of 1986 (SARA) and, to the extent practicable, the National Oil and Hazardous Substances Pollution Contingency Plan (NCP). This decision document explains the factual and legal basis tor selecting the remedy for this site. The information supporting this remedial action decision is contained in the administrative record for this site. The State of Michigan is expected to concur on the selected remedy.

Assessment of the site

Actual or threatened releases of hazardous substances from this site, if not addressed by implementing the response action selected in this Record of Decision (ROD), may present an imminent and substantial endangerment to public health, welfare, or the environment.

Description of the Selected Remedy

The selected remedy is the final remedy for the site. The remedy addresses the threats posed by principal threat wastes at the site. Principal threat wastes are defined as those source materials considered to be highly toxic or highly mobile that generally cannot be reliably contained or would present significant risk to human health or the environment should exposure occur.

The major components of the selected remedy include the following:

- Excavation and disposal of shallow and deep contaminated soil;
- Resampling of locations identified in the Remedial Investigation which showed contaminant levels above cleanup standards to determine the extent of contamination, and,
- Restoration of residential areas affected by excavation

Declaration of Statutory Determinations

The selected remedy is protective of human health and the environment, complies with federal and state requirements that are legally applicable or relevant and appropriate to the remedial action, and is cost effective. This remedy does not satisfy the statutory preference for remedies that reduce the toxicity, mobility, or volume as a principal element because treatment of the principal threats of the Site was not found to be practicable: it would not be cost effective to treat such a small volume of waste, and the residential nature of the site precludes on-site treatment. However, if the waste is found to be characteristically hazardous, it will be required to be treated prior to final disposal, and the remedy will then satisfy the preference for remedies that employ treatment that reduces toxicity, mobility, or volume as a principal element.

Because this remedy will not result in hazardous substances remaining on-site above health-based levels, the five year review will not apply to this action.

< IMG SRC 0596304A>

State of Michigan: Letter of Concurrence

< IMG SRC 0596304B>

August 26, 1996

Mr. William E. Muno, S-6J Director, Superfund Division U S. Environmental Protection Agency, Region 5 77 West Jackson Boulevard Chicago, Illinois 60604-3590

Dear Mr. Muno:

The Michigan Department of Environmental Quality (MDEQ), on behalf of the state of Michigan, has reviewed the proposed Record of Decision (ROD) dated June 19, 1996, for the Lower Ecorse Creek Superfund site in Wayne County, Michigan. We are pleased to inform you that we concur with the remedy outlined in the ROD.

This remedy meets state cleanup requirements, including the generic residential cleanup criteria, pursuant to Part 201 of the Natural Resources and Environmental Protection Act, 1994 PA 451, as amended (formerly known as the Michigan Environmental Response Act), and is expected to allow for unrestricted use of the site upon completion of the excavation.

The MDEQ looks forward to the successful completion of the final remedy for the Lower Ecorse Creek Superfund site. If you have any questions, please contact Mr. Brady Boyce, Superfund Section, Environmental Response Division, at 517-373-4824, or you may contact me.

Sincerely

Russell J. Harding Director 517-373-7917

cc: Mr. RMW, EPA

Mr. Alan J. Howard, MDEQ

Mr. Ardon Toland, MDEQ

 ${\tt Dr. George \ Carpenter, \ MDEQ}$

Mr. Brady Boyce, MDEQ/Lower Ecorse Creek File (GU)

Decision Summary for the Record of Decision

1.0 Site Name, Location, and Description

The Lower Ecorse Creek (LEC) Site is located in Section 17, R11E, T3SN in the City of Wyandotte, Wayne County, Michigan (Figure 1). The City of Wyandotte is located about 6 miles southwest of the City of Detroit. The site area includes six residential blocks centered around the 400 block of North Drive. The Detroit and Toledo Railroad tracks are located east of the residential area. The Ecorse River borders the site to the north and west. Directly north of the Ecorse River are the Dowriniver Communities Combined Sewer Overflow Treatment Plant and the abandoned Great Lakes Steel Foundry. Two lots located at 2303 Oak Street are also included as part of the site (Figure 2). The Oak Street Site is located approximately 1.5 miles west-northwest of the North Drive properties and the comer of 23rd Avenue and Oak Street.

2.0 Site History and Enforcement Activities

Before 1930, land near the banks of the Ecorse River in Wyandotte was wetlands. A 1937 aerial photograph shows the wetlands and a small brook that flowed to the river in the lots between Lots 23/24 (470/480 North Drive) and Lot 27 (446 North Drive). A 1951 photograph indicates that most of the wetland area had been filled, and residential development along North Drive had occurred.

By 1957 the river had been rechanneled. The confluence of the north and south branches of the river (also known as Upper and Lower Ecorse Creeks) was relocated from north of Lot 43 (304 North Drive). Extensive fill is evident north of the Ecorse River. Modifications to the river in the early 1980s involved straightening the south bank of the river at the rear of several residential properties in the area, reportedly using construction debris as fill. Interviews with local residents indicate that the homes on North Drive were built from about the 1920s through the 1980s.

In October 1989, the owner of the residence at 470/480 North Drive (Lots 23/24) reported to the Wayne County Health Department (WCHD) that workers excavating to replace the driveway on their property had encountered blue-colored soil. Preliminary tests by WCHD found high concentrations of cyanide in the blue-colored soil. WCHD consulted with the Agency for Toxic Substances and Disease Registry (ATSDR), and the agencies contacted U.S. EPA for further investigation. The U.S. EPA found that a large area of sod in the site area was colored a deep blue. The primary constituent of the coloring agent is ferric ferrocyanide. It is suspected that waste from a coal-gasification plant deposited in the site area is the probable source for the blue material found on the site. Blue-colored water was observed in the basement sump of the house at Lots 23/24. Blue stains also were seen on the basement walls of this house.

The ATSDR issued health consultations on the site in November 1989, July 1990, November 1990, and March 1991. In these consultations, ATSDR concluded that the site posed a significant health threat and recommended that residents avoid contact with contaminated areas until permanent measures could be completed.

In December 1989, the U.S. EPA covered the areas of visible contamination at the site with six inches of clean topsoil, to provide a temporary cover while further investigation went on and a permanent solution was developed. After it was reported that the new soil was eroding away, additional soil was added to the cover in August 1991. In January, 1993, the owner of the residence at Lots 23/24 reported that his basement had flooded with blue-colored water. U.S. EPA investigators found that these waters contained high concentrations of cyanide.

On August 13, 1993, the ATSDR issued a Public Health Advisory for the North Drive (Lower Ecorse Creek) site. The Advisory concluded that the levels of cyanide found in the soil at the site pose a significant public health hazard and that anyone using shallow groundwater in the site area may be at risk of exposure to cyanide contaminated water. The Advisory made the following recommendations:

1. Residents of the site area should be dissociated from the cyanide contamination;

- 2. Permanent remedial measures should be implemented as soon as possible;
- 3. The site should be considered for U.S. EPA's National Priorities List;
- 4. Residents in the area should be surveyed to locate any private wells in the site area;
- 5. Restrictions on digging in the site area should be considered; and,
- 6. The ATSDR Division of Health Studies should evaluate reports of adverse health effects to determine the source of these effects.

In November 1993, the U.S. EPA began a time-critical removal action at the site. This action included sampling 10 residential lots on North Drive for cyanide and other selected contaminants to delineate the area of contamination. Based upon those sampling results- contaminated soils from around the residence at Lots 23/24 and Lots 91/92 were removed and disposed of off-site. The foundations at both residences were also found to be deteriorated by the acidic nature of the waste. Repairs were made by U.S. EPA to both foundations. At Lots 23/24 application of a chemical resistant sealant to the basement walls and floors at the residence, and restoration of the surface drainage at the residence were also required. A U.S. EPA contractor had completed the excavation of contaminated soil around the residence and backfilled the area with clean soil by January 1994.

On January 19, 1994, the Lower Ecorse Creek site was proposed for listing on the National Priorities List (NPL) based upon the ATSDR Public Health Advisory. It became final on the NPL on May 31, 1994.

In March 1995, an area of apparent cyanide contamination, similar to the material identified at North Drive, was discovered at the residential lot at 2303 Oak Street in Wyandotte, Michigan. A time-critical removal action was initiated and contaminated soil was excavated and disposed of and the site was restored in May 1995 Because of the apparent sirn-arity of the material to that found at the LEC site, the Oak Street site was included in the remedial investigation and feasibility study for the LEC site.

On December 22, 1993 U.S. EPA issued General Notice letters to BASF Corporation and Michigan Consolidated Gas Company, offering them the opportunity to undertake the RI/FS for the site. Both parties refused to undertake the work and on March 14, 1994 a flind-financed RI/FS began. The final RI report was released to the public in February 1996. The final FS report was released to the public on April 15, 1996.

On November 29, 1995, ATSDR released a final Public Health Assessment for the site which stated that the recommendations made in the 1993 Public Health Advisory concerning the site have been addressed as follows:

- 1. Residents of the site area should be dissociated from the cyanide contamination—U.S. EPA carried out an Emergency Removal Action in Late 1993 and early 1994 at the site. Contaminated surface soil was removed and the walls and floor of the basement of the house at the center of the contaminated area were sealed to keep contaminated groundwater out;
- 2. Permanent remedial measures should be implemented as soon as possible The removal of contaminated sod in the 1993 U.S. EPA Emergency Removal Action is a permanent remedial measure. Permanent measures to remedy the contamination of the groundwater have not been implemented. Although groundwater is not used in the site vicinity, residents and visitors may be exposed to the water through seepage into basements in the site area;
- 3. The site should be considered for U.S. EPA's National Priorities List U.S. EPA placed the site on the NPL in January 1994;
- 4. Residents of the site area should be surveyed to locate private wells in the site area No private wells were identified in the immediate area surrounding the site. All households are connected to municipally supplied water;

- 5. Restrictions on digging in the site area should be considered U.S. EPA has advised residents of the site area not to dig in their yards; and,
- 6. The ATSDR Division of Health Studies should evaluate reports of adverse health effects to determine the source of these effects ATSDR and U.S. EPA evaluations of the health problems experienced by one young resident of the site area have not identified a connection between these effects and his potential exposure to the cyanide compounds in the soil, air, and groundwater at his home.

3.0 Highlights of Community Participation

The Responsiveness Summary in Section 12.0 discusses the involvement of the community during the RI/FS and remedy selection process and shows that the public participation requirements of CERCLA Sections 113(k)(2)(I-v) and 117 of CERCLA have been met at this site. The decision is based on the Administrative Record.

4.0 Scope and Role of Operable Unit-or Response-Action Within Site Strategy

This Record of Decision (ROD) addresses the final remedy for the site. The threats posed by this site to human health and the environment are primarily from cyanide contaminated soil. Other contaminants are present, e.g. SVOC's, however, they do not pose an unacceptable risk.

The contaminated soil is the source materials for contamination at the site and are classified as principal threat waste. Principal threat wastes are considered to be those source materials that are highly toxic or highly mobile that generally cannot be reliably contained or would present significant risk to human health or the environment should exposure occur.

4.1 Site Physical Characteristics

4.1.1 Topography

The LEC site is located in a former wetland area of the Ecorse River. The site consists of a developed residential area consisting of flat lying residential lots. Development of the residential area required filling in the former wetlands and later straightening the south bank of the Ecorse River.

4.1.2 Geology

Site geology primarily consists of fill soils and wetland and native fluvial deposits from the Ecorse River overlying lacustrine clays. Fill was historically used to develop wetland areas into a residential area along North Drive. The fill consists of construction debris, natural clay fill materials, and waste materials.

Native soils beneath the fill consist of gray to orange brown, finely layered, fine to very fine sand, silt, and clay. Native surficial soils appear to have been reworked, possibly during the construction of the residences.

4.1.3 Hydrology

The site is bounded on the north and west by the Ecorse River. The Ecorse River flows to the east and discharges into the Detroit River about 200 feet from the eastern site boundary. The area north of North Drive lies in the river's designated 100-year floodplain (see Figure 3).

Runoff from the northern portions of the residential lots on North Drive flows into the Ecorse River. The runoff from the remaining residential lots flows into the streets and storm sewers. The storm sewers carry the runoff to the local treatment plant, from where it is discharged into the river

4.1.4 Hydrogeology

Whether groundwater is present in clay rich terrain such as in eastern Wayne County depends on the occurrence of glaciofluvial deposits. Limited quantities of groundwater may be found in these permeable localized sand and gravel bodies that are buried within the lake plain deposits. The frequency and occurrence of these discontinuous sand and gravel bodies decreases toward the Detroit River (Mozola 1969).

Groundwater was not detected in most of the borings completed throughout the study area. Groundwater occurred in the borings only in thin permeable zones consisting of coarse fill debris and soft wetland soils. Primarily these localized isolated zones occurred within the fill near the river.

Deep borings drilled outside of fill areas for stratigraphic profiling indicated moist to wet soils only in wetland soils. This perched water yielded very small quantities of water and did not prove to be laterally extensive.

5.0 Summary of RI Findings and Previous Investigation Results

The U.S. EPA assigned CH2M HILL to perform an RI for the LEC and Oak Street sites. CH2M HILL developed and implemented an investigative approach that evaluated the nature and extent of contaminants in site soils, perched groundwater, surface water and sediment, and residential air, sump water, and sump sediment. Field activities were conducted from November 14 to December 22, 1994. Only surface and subsurface soils were investigated at the Oak Street properties. The findings of the RI conducted in the residential areas and those in the previous investigation area are summarized below. Tables 1 through 29 summarize the analytical results for each of the media tested. The area of previous investigation, is defined as those lots sampled during the removal action at this site (see Figure 1).

5.1 Nature and Extent of Contamination

5.1.1 LEC Soils

Background soil samples were collected in the residential areas adjacent to the LEC site and background concentrations were calculated according the MDEQ guidance document, Verification of Soil Remediation, 1994 (the mean plus 3 standard deviations). In the discussion, surface soils are assumed to be from 0 to 2-feet below ground surface and subsurface soils from 2 to approximately 17 feet below ground surface. This is consistent with the distinctions made between surface and subsurface soils to calculate risks in the baseline risk assessment.

In the area of the previous investigation, cyanide was detected in 73 percent of the surface soils at a maximum reported concentration of 1,730 mg/kg. Cyanide was also detected in the subsurface soils in most of the samples collected during the RI from the area of previous investigation at a maximum concentration of 32,300 mg/kg at a depth of 4 to 6 feet. The maximum cyanide concentration in the soil samples collected from the area outside the previous investigation area was 4.0 mg/kg, measured in a surface soil sample.

Antimony, barium, chromium, copper, iron, lead, manganese, and zinc were the metals detected most frequently at concentrations greater than background in both surface and subsurface soils at the site. Metals were detected above background most frequently -in the fill area adjacent to the Ecorse River.

As expected when analyzing for volatile organic compounds (VOCs) in a medium in close proximity to the open atmosphere, the majority of surface soils did not contain any detectable VOCs. Methylene chloride and acetone were the VOCs detected most frequently in surface soils at the site Methylene chloride was detected at a maximum concentration of 19:g/kg, and acetone was detected at a maximum concentration of 76:g/kg.

Methylene chloride, acetone, carbon disulfide, and 2-butanone were the VOCs detected in the subsurface soils with the greatest frequency at maximum concentrations of 260; 1,3005 65,000 and 196:g/kg, respectively.

Polynuclear aromatic hydrocarbons (PNAs) were detected across the site in both surface and subsurface soils. The maximum concentration of an individual PNA measured was 150,000 : g/kg for pyrene. PNAs were detected in the highest concentrations in the area of previous investigation and the northeast lot. Dioxin was found in soils at 10 site locations, primarily in the playground/park area at a maximum concentration of 16.0 ng/g. Two surface soil and two subsurface soil samples contained polychlorinated biphenyls.(PCBs). The maximum PCB

concentration was 250 :g/kg in a subsurface sod sample from the playground/park area.

Figure 4 shows the sampling locations where contaminants exceeded MDEQ cleanup standards.

5.1.2 Oak Street Site

At the Oak Street site, cyanide was detected in subsurface soils at concentrations ranging from 44.1 to 7,438 mg/kg. The highest concentrations were found the west area, adjacent to the asphalt parking lot. PAHs and other metals, similar to the ones detected at the North Drive area were also detected at the Oak Street Site.

5.1.3 Perched Groundwater

Low concentrations of VOCs and semivolatile organic compounds (SVOCs) (1 to 3 :g/kg) were detected in one perched groundwater sample from the playground/park area. Metals detected in the perched groundwater samples did not exceed background levels.

5.1.4 Surface Water and Sediment

Surface water and sediment samples were collected both upstream and adjacent to the Site. VOCs (1 to 15:g/L) were detected in four surface water samples. No SVOCs were detected in the surface water samples. Several inorganic analytes, including arsenic, barium, chromium, copper, lead, cyanide, zinc, and cadmium, were found in surface water samples. Low concentrations of acetone (less than 92:g/kg) and xylene (less than 10:g/kg) were identified in four sediment samples. SVOCs (19,760:g/kg total SVOCs) were detected at SD-02. Several metals, including cyanide, lead, and zinc, were detected in sediment at concentrations that exceed background sediment levels.

5.1.5 Residential Air, Sump Water, and Sump Sediment

No hydrogen cyanide was detected in the six air samples collected in the residential basements. Low concentrations of carbon disulfide and acetone were detected in one sump water sample (0.6 : g/L) and one sump sediment sample (14 : g/kg), respectively. Low concentrations of PNAs, phthalates, and phenols were detected in two sump water samples. Inorganic analytes, were detected in both sump water and sediment samples, with the highest concentrations found in the sample collected from the basement sump of the residence in the area of previous investigation.

5.1.6 Contaminant Fate and Transport

In general, contaminants in surface and subsurface soils have been identified during the RI in the highest concentrations in the area of previous investigation and in pockets of contaminated fill along the Ecorse River including the playground/park area, the northeast lot, and the north bank area (see Figure 1). The primary contaminant release and transport mechanisms from the LEC site consist of:

- Erosion, transport, and deposition of contaminated dust by wind
- Leaching of dissolved contaminants into perched groundwater and transport in groundwater to discharge areas such as the Ecorse River or potentially into residential basements by seepage through basement walls
- Surface runoff of dissolved contaminants to the Ecorse River or by soil erosion and particulate transport in surface water
- Volatilization of VOCs from the soil and migration offsite through the atmosphere, and possibly into basements.

The main contaminants at the site, Including PNAs, cyanide in the form of ferric ferrocyanide, and metals, tend to be persistent in the environment because they (PNAs) are slow to degrade and have low mobility.

Contaminants at the site are not expected to migrate a great distance from the source areas. Because there does not appear to be continuous groundwater unit at the site and because the perched groundwater identified is of limited aerial extent and depth, the groundwater pathway for contaminant migration is not considered to be significant. The primary migration pathways at the site are through the air and surface water run-off.

6.0 Risk Assessment

Pursuant to the National Contingency Plan (NCP) a baseline risk assessment was performed using analytical data generated during the RI and the removal project. The baseline risk assessment assumes no corrective action will take place and that no site-use restrictions or institutional controls such as fencing, groundwater use restrictions or construction restrictions will be imposed. However, for the future site scenarios, present action at the site and current plans for development are considered. The risk assessment determines actual or potential carcinogenic risks and/or toxic effects the chemical contaminants at the site pose under current and future land use assumptions using a four step process. The four step process includes: contaminant identification, health effects assessment, exposure assessment and risk assessment. Table 30 summarizes the results of the risk assessment.

6.1 Contaminant Identification

During the RI several chemicals in different media were detected and a list of "chemicals of potential concern" was developed using the following criteria:

- Any chemical detected at least once in any on-site soil, groundwater, surface water, or sediment sample was considered to be a possible chemical of concern;
- Several chemicals known to be essential for human nutrition were eliminated. These chemicals were present at levels that are considered non-toxic.
- Compounds that were detected at concentrations above the calculated background concentrations were retained as compounds of concern. According to RAGS Part A, most organic compounds found at remediation sites are not naturally-occurring, and thus cannot be eliminated from the quantitative assessment. The organic compounds detected were retained as compounds of concern.

The chemicals of potential concern are listed in Table 31.

6.2 Human Health Effects

The health effects for the contaminants of concern may be found in the RI report.

6.3 Exposure Assessment

The baseline risk assessment examined potential pathways of concern to human health under both current and future land-use scenarios for the immediate property and the surrounding area. The exposure scenarios which were evaluated in the baseline risk assessment were based on the residential land use that currently exists in the study area. It was also assumed also that the area would remain residential in the future. The residential land use scenario provided a conservative estimate of intakes, and therefore, risks.

The following pathways were selected for detailed evaluation under both the current and future land-use conditions:

- Residential adult and child exposure to surface and subsurface soil
- · Residential adult and child exposure to surface water and sediment in the Ecorse River
- Adult exposure to sump water and sump sediment

The current and future site uses are expected to be residential. Exposure to soil was evaluated separately

for the residential area and the playground/park area because the exposure assumptions for each of these areas were different. In addition, the risk to residents in the residential area was also evaluated separately for the area of previous investigation (for cyanide) and the rest of the residential area. Potential, human health impacts from exposure to surface soil, subsurface soil, surface water and sediments, and sump water and sediments are presented below.

6.4 Risk Characterization

For each potential human receptor, site-specific contaminants from all of the relevant routes of exposure were evaluated. Both non-carcinogenic and carcinogenic health effects were estimated.

Reference doses (RfDs) have been developed by U.S.EPA as a means of identifying the potential for adverse health effects from exposure to chemicals that typically exhibit non-carcinogenic effects. RfDs, which are expressed in units of mg/kg-day, are estimates of average daily exposure levels for humans, including sensitive individuals. Estimated intakes of chemicals from environmental media (e.g., the amount of chemical ingested from contaminated drinking water) can be compared to the RfD. RfDs are derived from human epidemiological studies or animal studies to which uncertainty factors have been applied (e,g., to account for the use of animal data to predict effects on humans). These uncertainty factors help ensure that the RfDs will not underestimate the potential for adverse non-carcinogenic effects to occur.

The Hazard Index (HI), an expression of non-carcinogenic toxic effects, measures whether a person is being exposed to adverse levels of non-carcinogens. The HI provides a useful reference point for gauging the potential significance of multiple contaminant exposures within a single medium or across multiple media. The HI for non-carcinogenic health risks is the sum of all contaminants for a given scenario. Any Hazard Index value greater than 1.0 suggests that a non-carcinogen potentially presents an unacceptable health risk.

Cancer potency factors (CPFs) have been developed by EPA's Carcinogenic Assessment Group for estimating excess lifetime cancer risks associated with exposure to potentially carcinogenic chemicals. CPFs, which are expressed in units of (mg/kg-day) -1, are multiplied by the estimated intake of a potential carcinogen, in mg/kg-day, to provide an upper-bound estimate of the excess lifetime cancer risk associated with exposure at that intake level. The term "upper bound" reflects the conservative estimate of the risks calculated from the CPF. Use of this approach makes underestimation of the actual cancer risk highly unlikely. Cancer potency factors are derived from the results of human epidemiological studies or chronic animal bioassay. The excess lifetime cancer risks are the sum of all excess cancer lifetime risks for all contaminants for a given scenario determined by multiplying the intake level by the cancer potency factor for each contaminant of concern.

6.4.1 Surface Soils

Potential inadvertent Ingestion and dermal absorption of contaminants detected in surface soils in the residential area resulted in estimated potential excess lifetime cancer risks of 9 x 10 -6 for reasonable maximum exposure (RME) adults and 1 x 10 -6 for RME children. The central tendencies exposure (CTE) cancer risk were estimated at 2 x 10 -7 for adults and 1 x 10 -6 for children. These cancer risks were due primarily to carcinogenic PNAs (Class B2 carcinogen) detected in residential area surface soils. Inhalation of VOCs and particulates resulted in estimated potential lifetime cancer risks of less than 1 x 10 -6 for both adults and children. Hazard indexes (HIs) due to exposure to surface soils for adults and children were less than 1.

The HIs for both adults and children with exposure to cyanide-containing surface soils from the area of previous investigation were less than 1. Because the adult and child RME His were less than 1, the CTE scenario was not evaluated.

Potential inadvertent ingestion of and dermal contact with contaminants detected in surface soil samples from the playground/park area resulted in estimated potential excess lifetime cancer risks of 1 x 10 -5 for RME adults and 3 x 10' for RME children. The estimated potential excess lifetime cancer risk for CTE adults was 6 x 10 -7 , and the excess lifetime cancer risk for CTE children was 5 x 10 -6 These cancer risks estimates were due primarily to the presence of arsenic (Class A carcinogen) in playground/park surface soils. Inadvertent ingestion of surface soils resulted in potential excess lifetime cancer risks of 6 x 10 -6 for

RME adults and 1 x 10 -5 for RME children; the excess lifetime cancer risk estimated for CTEs due to ingestion of soil was 4 x 10 -7 for the adult, and 5 x 10 -6 for the child. The majority of this risk was due to the presence of arsenic and PNAs. Dermal absorption of surface soil contaminants from the playground/park area also contributed substantially to the estimated cancer risks, with adult dermal exposure resulting in a cancer risk of 5 x 10 -6 (RME) and child dermal exposure resulting in a cancer risk of 2 x 10 -5 (RME). The cancer risk estimated for CTEs due to dermal contact was 2 x 10 -7 for the adult and 2 x 10 -7 for the child. Arsenic, PCBs, and PNAs in playground/park surface soils were the primary carcinogens for these risk estimates. Inhalation of surface soil contaminants resulted in estimated potential lifetime cancer risks well below 1 x 10 -6. HIs associated with exposure to playground/park surface soils were less than 1 for adults and children.

6.4.2 Subsurface Soils

Future potential inadvertent ingestion of contaminants detected in residential area subsurface soils resulted in estimated potential excess lifetime cancer risks of 5 x 10 -6 for RME adults and 1 x 10 -5 for RME children. The estimated potential excess lifetime cancer risk for CTE adults was 2 x 10 -7, and the estimated potential excess lifetime cancer risk for CTE children was 4 x 10 -6. The cancer risk was due to the presence of PNAs. This risk was caused by the ingestion of contaminated subsurface soils; the risk due to inhalation of contaminated subsurface soils was below 1 x 10 -6 Dermal risks for PNAs, although not calculated, should be assumed to be of the same order of magnitude as those calculated for ingestion. HIs for subsurface soil exposures were less than 1 for adults and children.

The presence of cyanide at a concentration of 32,300 mg/kg in subsurface soils in the area of previous investigation (hot-spot) resulted in an HI estimate greater than 1 for RME children, indicating the potential for noncancer adverse health impacts. The RME HI was due to ingestion of contaminated subsurface soils in this area (2.1 for children), with dermal exposure contributing a much smaller portion (0.2 for children). The HI for RME adults was less than 1. The CTE adult and child HIs were less than 1.

6.4.3 Surface Water and Sediment

Exposure to surface water and sediments resulted in estimated potential excess lifetime cancer risks below 1×10^{-6} , HIs. associated with ingestion and dermal exposures were less than 1. Because RME risks were below 1×10^{-6} (and HIs were less than 1), the CTE scenario was not evaluated.

6.4.4 Sump Water and Sediments

Dermal contact with sump water by adults resulted in an estimated potential excess lifetime cancer risk below 1×10^{-6} , as did contact with sump sediments. The HIs associated with dermal contact with sump water and sump sediments were both less than 1. Because RME risks were below 1×10^{-6} (and HIs were less than 1), the CTE scenario was not evaluated In an assessment of the risks to household pets from ingestion of sump water there were no unacceptable risks found. For a full explanation of the household pet risk evaluation please see the June 13, 1996, memorandum from CH2M Hill, which is in the administrative record. The risk to a household cat was evaluated assuming both a one year and 14 year exposure duration with a consumption of 0.3 liter/day. A comparison of the intake and the oral toxicity value for each chemical of concern showed no unacceptable risk.

6.4.5 Acidic and Basic Soils

Due to the nature of soils in the area, a qualitative evaluation of the risks due to exposure to acidic or basic soils was also conducted. Acidic or basic soils at the site are found at the site in the areas of high cyanide concentrations. Exposure to acids can result in severe skin burns, usually with a dry crust from coagulation necrosis. Alkalies (bases) produce softer burns, which can be extremely painful. Acid bums to the eye are a dual function of the pH and the capacity of the acid's anion to combine with ocular proteins (in addition to other aspects, such as the defatting action by sulfuric acid and sulfur dioxide). The effects noted above for exposure to acidic or basic solutions may or may not occur upon exposure to soils. The effects may vary in seventy depending on the matrix of the acid or base (solution vs. soil) and the physical condition of the skin or eye (thickness, presence of cuts or abrasions, etc.). Aside from the acidic or basic

nature of soil, the physical abrasive action of soils would damage the skin or the eyes. The chemical action of acidic or basic soils is likely to have a greater impact on physical structures, such as foundations, due to their continuous and long-term contact with the soils and weathering effects (freeze-thaw).

6.4.6 Ecological Risk Assessment (ERA)

The purpose of the ERA is to evaluate the potential adverse ecological effects that may be or are occurring as a result of exposure to site-related stressors at the LEC site. The ERA evaluates potential threats to ecological receptors in the absence of any remedial actions.

6.4.6.1 Aquatic Communities

The greatest risk posed by contaminants associated with the LEC Site appears to be from contaminants within sediments of Lower Ecorse Creek. However, the contaminants associated with the site and found in the sediments were also detected upstream from the site indicating that the site is not the source of the contamination problem in the creek. The habitat associated with the creek itself already precludes the existence of a diverse and sustainable population of aquatic organisms. Even though the habitat quality is questionable, exposure and risk to aquatic organisms was evaluated by comparing exposure dose estimates to National Ambient Water Quality Criteria (NAWQC)standards or literature-based benchmark value. Aluminum, barium, lead, iron, anthracene, and benzo(a)anthracene are present and are of the greatest concern to aquatic organisms based on the dose estimate comparisons. The NAWQC and sediment benchmark values were considered conservative. Therefore, the actual affect on the aquatic ecosystem may not be as great as indicated by the dose estimate comparisons.

6.4.6.2 Terrestrial Communities

The contaminants associated with surface soils that pose a potential threat to terrestrial communities associated with the LEC Site were lead and cyanide. Even though the residential setting does not support a diverse community structure, the level of cyanide in the soil may be of significant threat to birds such as robins. It should be noted here that several COCs were not evaluated relative to effects on terrestrial communities because there is limited toxicological information available.

6.5 Rationale for Further Action

Actual or threatened releases of hazardous substances from this site, if not addressed by implementation of the response action selected by this ROD, may present an imminent and substantial endangerment to the public health, welfare or the environment. Therefore, based upon the findings of the RI report and the discussion above, a Feasibility Study (FS) was performed to focus the development of alternatives to address the threats at the site. The FS report documents the evaluation of the magnitude of the site risks, site-specific applicable or relevant and appropriate requirements, and the requirements of CERCLA and the NCP in the derivation of remedial alternatives for the LEC site.

7.0 Description of Alternatives

Three alternatives for the remediation of soils at the LEC site were developed including a no action alternative. These alternatives include all the remedial technologies remaining after screening that are applicable to inorganic and SVOC contamination. The alternatives are:

Alternative 1-No action

Alternative 2-Excavation and disposal of shallow contaminated soil and implementation of institutional controls for areas of deep contaminated soil

Alternative 3-Excavation and offsite treatment and disposal of shallow and deep contaminated soil

7.1 Alternative 1-No Action

Capital Cost: None
Annual Operation and Maintenance Cost None
Present Worth None
Time to Implement None

The no action alternative is required by the NCP. Its purpose is to allow comparison of alternatives to the conditions that currently exist and that would exist in the future. Under Alternative 1 there would be no remediation of the contaminated soils. There would be no ongoing site security or installation of a site fence. No restrictions would be placed on sale of the property or future development of the site.

7.2 Alternative 2-Excavation and Disposal of Shallow Contaminated Soils and Institutional Controls for Deep Contaminated Soils

Capital Cost \$894,150
Operation and Maintenance Cost None
Present Worth \$894,150

Time to Implement 6 months from start of construction

The major components of Alternative 2 are:

A Excavation of shallow contaminated soil

A Disposal of shallow contaminated soil

A Implementation of institutional controls for deep contaminated soil

A Restoration of residential areas

The objective of Alternative 2 is to protect human health and the environment from unacceptable risk associated with direct contact with the soils through the use of a combination of excavation, disposal, and deed restrictions.

At the location where contaminated soils were detected in the 0 - 2 foot depth range, those locations would be excavated to 1 foot below the level of contaminated soil using standard excavation equipment such as backhoes, front-end loaders, and bulldozers. Since the soils to be excavated are located on the property of private residences, small excavation equipment and tools for hand digging will also be required. Trucks used for loading of the excavated soils would be direct-loaded, and stockpiling of soil would be minimized. Excavation would proceed downward and outward from the centers of the known areas of contamination. Assuming an area of 10 feet by 10 feet and a depth of 2 foot deeper than the level of contamination, the amount of inorganic- and SVOC -contaminated shallow soil to be excavated and disposed of offsite at a landfill is estimated to be 298 cubic yards. This total is an estimate and is subject to increase if confirmatory sampling indicates that additional soil removal is necessary. Further sampling will be performed in areas that have been designated for remediation and which are located on private property and appear to be isolated areas of contamination. This additional sampling will confirm the RI sampling results and delineate areas of contaminated soil so that property owners will know if excavation is required and, if so, how much of their property will be by excavation before remedial activities on their property begin.

7.3 Alternative 3-Excavation and Disposal of Shallow and Deep Soil

Capital Cost \$645,800
Annual Operation and Maintenance Cost None
Present Worth \$645,800

Time to Implement 3 months from start of construction

The major components of Alternative 3 are:

- Delineation of isolated contaminated soil on private property
- Excavation of shallow and deep contaminated soil
- Disposal of shallow and deep contaminated soil
- Restoration of residential areas

Alternative 3 would eliminate the need for institutional controls because all contaminated soils with concentrations of COCs above the remedial goals established for the site, and the unacceptable health risks associated with them, would be removed from the site. Shallow and deep contaminated soils would be removed using procedures discussed for Alternative 2. At locations where R1 data indicate that only the deep soils are contaminated, soils above the zone of contamination would be stockpiled in a clean area to be used later as backfill after confirmatory sampling is completed. Contaminated soil would be excavated. The amount of inorganic- and SVOC-contaminated soil to be excavated and disposed of offsite at a landfill is estimated to be 906 cubic yards. The total is an estimate, confirmatory sampling will be necessary to determine the actual extent of contamination.

Further sampling will be performed in areas that have been designated for remediation and which are located on private property and appear to be isolated areas of contamination. This additional sampling will confirm the RI sampling results and delinate areas of contaminated soil so that property owners will know if excavation is required and, if so, how much of their property will be impacted by excavation before remedial activities on their property begin.

8.0 Summary of Comparative Analysis of Alternatives

The relative performance of each remedial alternative was evaluated in the FS and below using the nine criteria set forth in Title 40 of the Code of Federal Regulations (40 CFR) Section 300.430 of the NCP An alternative providing the "best balance" of trade-offs with respect to the nine criteria is determined from this evaluation.

8.1 Threshold Criteria

The following two threshold criteria, overall protection of human health and the environment, and compliance with Applicable or Relevant and Appropriate Requirements (ARARs) are criteria that must be met in order for an alternative to be selected.

8.1.1 Overall Protection of Human Health and the Environment

Overall protection of human health and the environment addresses whether a remedy eliminates, reduces, or controls threats to human health and to the environment.

Alternative 1, no action, does not provide overall protection to human health and the environment. Direct contact with contaminated soils can still occur.

Alternatives 2 and 3 protect human health and the environment by removing or treating contaminated soil from the LEC and Oak Street sites. These alternatives, however, increase short-term risks onsite during the excavation and consolidation of soil and the protectiveness offsite would rely on the integrity of the offsite landfill.

Alternative 3 protects human health and the environment by removing and treating contaminated soils. Alternative 2 also protects human health and the environment to some degree by removal of shallow contaminated soil; however, this alternative provides no protection from the deep soils with concentrations exceeding cleanup standards, therefore, institutional controls would be required. The protectiveness provided by the offsite disposal facility would rely on the effectiveness of the stabilization/solidification process. Short-term risks are elevated onsite for workers and the community during the excavation and treatment of contaminated soils.

Alternatives 2 and 3 are both protective of human health and the environment because direct exposure from contaminated soil would be prevented as long as exposure to deep contaminated soil is prevented under Alternative 2.

8.1.2 Compliance with Applicable or Relevant and Appropriate Requirements

Alternative 1 does not comply with federal or state promulgated standards such as those specified in Michigan Act 451, Part 201. Alternative 2 would not reduce the risks from the deep contaminated soils that were identified in the risk assessment and the release and transport mechanisms would remain unchanged. Alternative 3 would reduce site-related risks and would comply with all ARARs if the required permits are obtained.

8.2 Primary Balancing Criteria

8.2.1 Long-Term Effectiveness and Permanence

This criterion refers to the expected residual risk and the ability of an alternative to maintain reliable protection of human health and the environment over time once clean up levels have been met.

Alternative 1, no action, would not provide long-term effectiveness. Exposure and risks resulting from current conditions at the site would continue.

Alternatives 2 and 3 provide long-term effectiveness at the site by permanently removing or treating contaminated soils that are above cleanup standards. Alternative 2 does not provide as much long-term protectiveness because the deep contaminated soils would still be in place. The SVOC COCs at the LEC site are, however, relatively immobile. The contaminated sod disposed of offsite would be controlled by measures taken at the disposal facilities. With present regulations on designing, constructing, and operating disposal facilities, long-term effectiveness would be expected.

Alternatives 2 and 3 are protective, however, Alternative 3 is more protective because both the shallow and deep soil are removed and additionally controlled by the stabilization/solidification process. Inorganic chemicals can be stabilized in the long term because they are chemically stabilized within the cement and the SVOCs are physically solidified within the mass of concrete, reducing the mobility of the SVOCs. Although soils contaminated with SVOCs and inorganic cherruicals would remain onsite in Alternative 2, the potential for direct exposure would be reduced by the enforcement of deed restrictions. These are meant to prevent activities that involve exposure of the deep contaminated soils.

8.2.2 Reduction of Toxicity, Mobility, or Volume through Treatment

In Alternative 1, no action, toxicity and volume of the contaminants would not change. While the mobility of the COCs will not change, they may be transported to uncontaminated areas by water erosion of the soil and infiltration.

In Alternatives 2 and 3, while there are no reductions in the toxicity and mobility of the COCs, risks to the community associated with the toxicity and mobility of the COCs are reduced when the contaminated material is disposed of or treated offsite. The offsite disposal or treatment facility would have measures in place to control the toxicity and mobility of the COCs brought from the site. Both Alternatives 2 and 3 are effective at reducing migration through the stabilization/solidification of excavated soils but would increase the volume of the excavated soil by approximately 120 percent by the addition of the stabilizing material. This increase in the volume of soil would not impact the site, however, since the stabilization/solidification process would be conducted offsite at the disposal facility.

8.2.3 Short-Term Effectiveness

Under Alternative 1, there would be no additional short-term risk to the community. Alternatives 2 and 3 create greater short-term risks to the community and onsite workers due to the excavation and disposal/treatment of contaminated soils. Alternative 2 will not create as much risk since only the shallow

contaminated soil will be removed. Short-term risks can be reduced through common construction practices, such as using Level D personal protection. Compared to no action, Alternatives 2 and 3 create some potential for direct contact by residents living in the houses in the immediate vicinity of the work, with contaminated soil during excavation and disposal.

Some additional considerations are the noise generated and typical risks associated with the heavy equipment onsite. Heavy traffic caused by vehicles used for excavation and trucks for soil disposal and site restoration will also be associated with all Alternatives, except for the no action, alternative. Added risks are associated with this traffic since it will be in the midst of a residential area. RAs associated with Alternative 3 can be performed within 3 months of the start of construction; Alternative 2 may take longer since there may be time delays in obtaining deed restrictions for the individual residential properties.

8.2.4 Implementability

This criterion addresses the technical and administrative feasibility of implementing an alternative, and the availability of vanious services and materials required for it implementation.

The technical feasibility of all alternatives is well understood. The technologies associated with Alternatives 2 and 3 are straightforward and usually easy to implement, however, because the areas to be excavated are located in a residential area, implementation becomes complicated. Applying deed restrictions for Alternative 2 may be administratively difficult to implement.

Alternatives 2 and 3 are both technically feasible.

8.2.5 Cost

This criterion compares the capital, operation and maintenance (O&M), and present worth costs for implementing the alternatives at the site. Table 32 shows the cost summary.

| Table 32 | | | | | | | | | |
|-------------|------|---------|--|--|--|--|--|--|--|
| Alternative | Cost | Summary | | | | | | | |

| | Capital Cost | O&M Cost | Present Worth |
|---------------|--------------|----------|---------------|
| Alternative 1 | None | None | None |
| Alternative 2 | \$894,150 | None | \$894,150 |
| Alternative 3 | \$645,800 | None | \$645,800 |

8.3 Modifying Criteria

8.3.1 State Acceptance

State acceptance indicates whether, based on its review of the RI/FS, and Proposed Plan, the State concurs with, opposes, or has no comment on the preferred alternative.

The State of Michigan has assisted in the development and review of the Administrative Record. The State's position regarding the selected alternative will be presented in a formal statement of concurrence, if appropriate. The State is expected to concur on the remedy.

8.3.2 Community Acceptance

The specific public comments received and U.S. EPA's responses are outlined in the attached Responsiveness Summary.

9.0 The Selected Remedy

Based upon considerations of the requirements of CERCLA, the NCP, and balancing of the nine criteria, the U.S. EPA has determined that Alternative 3, Excavation and Disposal of Shallow and Deep Contaminated Soil, is the most appropriate remedy for the site. The components of the selected remedy are described below.

The major components of the selected remedy are:

Delineation of isolated areas of contaminated soil on residential properties

Excavation of shallow and deep contaminated soil

Off-site disposal of shallow and deep contaminated soil

Restoration of properties affected by the remediation

9.1 Delineation of Isolated Contaminated Soil on Private Property

Further sampling will be performed in areas that have been designated for remediation, are located on private property and appear to be isolated areas of contan-driation. This additional sampling will confirm the RI sampling results and delineate areas of contaminated soil so that property owners will know if excavation is required and, if so, how much of their property will be impacted by excavation before remedial activities on their property begin.

9.2 Excavation of Contaminated Soil

Shallow and deep soils, with contaminants above the cleanup standards set forth in Table 33 would be excavated and disposed of off-site at an approved disposal facility. At locations where RI data indicate that only the deep soils are contaminated, soils above the zone of contamination would be stockpiled in a clean area to be used later as backfill after confirmatory sampling is completed. Contaminated soil would be excavated.

Upon removal of the contaminated soil, soil samples would be collected from the bottom and perimeter of the excavation to verify that soils with concentration levels above the PRGs have been removed. Following excavation, the hole would be backfilled using uncontaminated soil.

9.3 Disposal at a Landfill

The excavated soil contaminated above the cleanup standards would be transported to a Subtitle D landfill for disposal. Testing for RCRA hazardous waste characteristics would be performed, and if necessary, the soil would be stabilized off-site to meet the disposal facility's requirements prior to landfilling.

9.4 Restoration of Residential Areas

Excavation around private residences could require removal of sidewalks, lawns and other vegetation. Areas affected by the remediation will be restored, as close as practicable to their existing condition (including trees and shrubs).

10.0 Statutory Determinations

U.S. EPA's primary responsibility at Superfund Sites is to undertake remedial actions that protect human health and the environment. Section 121 of CERCLA has established several statutory requirements and preferences. These include the requirement that the selected remedy, when completed, must comply with all applicable, relevant and appropriate requirements ("ARARS") of Federal and State environmental laws, unless the invocation of a waiver is justified. The selected remedy must also provide overall effectiveness appropriate to its costs, and use permanent solutions and alternative treatment technologies, or resource recovery technologies, to the maximum extent practicable. Finally, the statute establishes a preference for remedies which employ treatment that significantly reduces the toxicity, mobility or volume of contaminants.

10.1 Protection of Human Health and the Environment

Implementation of the selected remedy will protect human health and the environment by reducing the risk of exposure to hazardous substances present in surface soils and subsurface soils at the Site. Excavation and

off-site disposal of the contaminated soil will minimize the direct contact risk of exposure to hazardous substances present in soil at the Site. Additionally, it will minimize the risk of drainage waters carrying the contaminants, via the drainage systems, or cracks in the foundations into the basements of the homes on site. It will also minimize the possibility of the acidic or basic soils associated with the contamination from coming in contact with and damaging foundation walls or utility lines. No unacceptable short-term risks Will be caused by implementation of the remedy. The community and site workers may be exposed to dust and noise nuisances during excavation, however, mitigative measures will be taken during remedy construction activities to minimize such impacts of construction upon the surrounding community and environs. Ambient air monitoring will be conducted and appropriate safety measures will be taken if contaminants are emitted.

10.2 Compliance with ARARs

The selected remedy will comply with all identified applicable or relevant and appropriate federal requirements, and with those state requirements which are more stringent, unless a waiver is invoked pursuant to Section 121 (d)(4)(B) of CERCLA.

Section 121 (d) of CERCLA requires that remedial actions meet legally applicable or relevant and appropriate requirements (ARARs) of other environmental laws. Legally, "applicable" requirements are those cleanup standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations promulgated under Federal or State law that specifically address a hazardous substance, pollutant, contaminant, remedial action, location, or other circumstances at a CERCLA site. "Relevant and appropriate" requirements are those requirements that, while not legally applicable to the remedial action, address problems or situations sufficiently similar to those encountered at the site that their use is well suited to the remedial action.

Non-promulgated advisories or guidance documents issued by federal or state governments ("to-be-considered or TBCs") do not have the status of ARARs; however, where no applicable or relevant and appropriate requirements exist, or for some reason may not be sufficiently protective, non-promulgated advisories or guidance documents may be considered in determining the necessary level of cleanup for protection of human health and the environment.

For a complete list of ARARs and TBCs for the alternatives at this site, see the FS Report. Below is a discussion of the key ARARs for the selected remedy.

10.2.1 Chemical-specific-ARARs

Chemical-specific ARARs regulate the release to the environment of specific substances having certain chemical characteristics. Chemical-specific ARARs typically determine the extent of clean-up at a site.

10.2.1.2 Federal ARARs

Clean Air Act National Ambient Air Quality Standards 40 CFR 50 - These regulations provide air emission requirements for actions which may release contaminants into the air. As the selected remedy involves excavation activities which may release contaminants or particulates into the air, emission requirements pomulgated under this act are relevant and appropriate.

10.2.1.2 State ARARs

Michigan Natural Resources and Environmental Protection Act 451, Part 201, as amended (formerly Act 307) - These regulations provide guidelines and cleanup standards for contaminated sites based on background levels, detection limits, health-based criteria, and current or anticipated land use, U.S. EPA's selected soil cleanup standards for this site are in compliance with Act 451, Part 201 and its implementing rules in that they meet the standard for selection of standards. The cleanup levels for contaminants in soil are determined by comparing current concentrations of contaminants with background concentrations and with allowable concentrations based on (1) risks and (2) ARARs. Table 33 lists the representative chemicals found in soil and the corresponding cleanup standards. These regulations are considered applicable to the Site.

Michigan Natural Resources and Environmental Protection Act and Air Pollution Rules, Michigan Ambient Air Quality Standards (MAAQS), Act 451, Part 55 (formerly Act 348 of 1965) - This act provides air emission requirements for actions which may release contaminants into the air. The selected remedy involves excavation activities which may release contaminants or particulates into the air. This act is relevant and appropriate.

10.2.2.0 Location-specific ARARs

Location-specific ARARs are those requirements that relate to the geographical position of the site. These include:

10.2.2.1 Federal ARARs

Clean Water Act Section 404 - This section of the Act regulates the discharge of dredge and fill materials at sites to waters of the United States. These regulations are applicable to excavation and backfilling activities which may take place adjacent to the creek in the former wetlands area.

Floodplain Management Executive Order 11988 - This order is applicable at this site. It requires the minimization of potential harm to or within flood plains and the avoidance of long and short term adverse impacts associated with the occupancy and modification of flood plains.

10.2.2.2 State ARARs

Soil Erosion and Sedimentation Control Act, Act 347 of 1972 - This act is applicable to this site because of the selected remedy's use of construction activities that may impact Ecorse Creek. The act regulates earth changes, including cut and fill activities which may contribute to soil erosion and sedimentation of surface water of the State. Act 347 would apply to any such action where more than one acre of land is affected or regulated action occurs within 500 feet of a lake or stream.

10.2.3.0 Action-specific ARARs

Action-specific ARARs are requirements that define acceptable treatment and disposal procedures for hazardous substances. These include:

10.2.3.1 Federal ARARs

RCRA Subtitle C Land Disposal Restrictions (LDRs) - These regulations govern the storage and disposal of hazardous waste. These regulations will be applicable to any site generated wastes which are characterized as hazardous waste. Soils at the site may exceed the TCLP standard for characteristic waste. If so, treatment of the excavated soils would be required prior to disposal in an off-site landfill.

10.2.3.2 State ARARs

Inland Lakes and Streams Act, Public Act 346 of 1972, as amended - The act regulates construction activities on or above bottomlands of inland lakes and streams. This act will be applicable to the selected remedy, because it addresses the mitigation of potential run-off, erosion, silting, and sedimentation in surface waters during construction.

Michigan Natural Resources and Environmental Protection Act, Act 451, Part 17 (formerly Act 127 of 1970) - This act prohibits any action that pollutes, impairs, or destroys the natural resources of the State. This act is applicable to the site since the site and the Ecorse Creek are viewed as natural resources.

Michigan Natural Resources and Environmental Protection Act, Act 451 Part 111 of 1994 (formerly Act 64 of 1979), as amended - This act regulates the generation, transportation, treatment, storage and disposal of hazardous waste. This ARAR will be applicable if the waste being disposed of off-site is characterized as hazardous.

Michigan Natural Resources and Environmental Protection Act, Act 451, Part 115 of 1994 (formerly Act 641 of

1978), as amended - This act regulates the disposal of non-hazardous solid waste. This act will be applicable for the off-site disposal of any waste which is non-hazardous.

10.3 Cost Effectiveness

Cost effectiveness compares the effectiveness of an alternative in proportion to its cost of providing environmental benefits. It is estimated that the cost of implementing the selected remedy will be \$646,000 in total capital costs. There are no costs associated with operation and maintenance. Appendix B, Table B-3 contains the detailed cost estimate for the selected remedy.

The selected remedy, Alternative 3, has been determined to afford overall effectiveness proportional to its cost. Alternative 3 carries moderate costs in comparison to the other alternatives considered. The No Action alternative, the alternative less costly than Alternative 3, does not offer the protectiveness provided by Alternative 3, because it leaves contaminants on-site.

10.4 Utilization of Permanent Solutions and Alternative Treatment Technologies or Resource Recovery Technologies to the Maximum Extent Practicable

The selected remedy represents the maximum extent to which permanent solutions and treatment technologies can be used in a cost-effective manner at this site. Of those alternatives that are protective of human health and the environment and that comply with ARARS, U.S. EPA has determined that the selected remedy provides the best balance in terms of long-term effectiveness and permanence, reduction of toxicity, mobility, or volume of contaminants, short term effectiveness, implementability, and cost, taking into consideration State and community acceptance.

The excavation and off-site disposal of the shallow and deep contaminated soil will provide the most permanent solution practical, proportionate to the cost.

10.5 Preference for Treatment as a Principal Element

Based on current information, U.S. EPA and the State of Michigan believe that the selected remedy is protective of human health and the environment and utilizes permanent solutions and alternative treatment technologies to the maximum extent possible. The remedy, however, does not satisfy the statutory preference for treatment of the hazardous substances present at the site as a principal element because such treatment was not found to be practical or cost effective. If characterized as hazardous, however, the material will be treated off-site prior to disposal.

11.0 Summary

The selected remedy will satisfy the statutory requirements established in Section 121 of CERCLA, as amended by SARA, to protect human health and the environment, will comply with ARARs, will provide overall effectiveness appropriate to its costs, and will use permanent solutions and alternate treatment technologies to the maximum extent practicable.

FIGURES

- < IMG SRC 0596304D>
- < IMG SRC 0596304E>
- < IMG SRC 0596304F>
- < IMG SRC 0596304G>

TABLE 1

TABLE
SURFACE SOIL INORGANIC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

Maximum Site Specific MDNR Residential Minimum Total Positive Detection Detected Detected Background Cleanup Criteria Positive Detections Detection Frequency Parameter Concentrations Direct Contact Exceeding Background Exceeding Background Mean UCL 95% Analyses Detections Frequency Value Value 0 Aluminum 43 43 100% 3150 14800 25958 JD 0 % 9001.73 10005.26 mg/kg Antimony 42 38 90% 0.16 7.0 0.45 150 12 29% 0.65 0.90 mg/kg Arsenic 165 164 99% 1.30 24.7 13.6 5.5 3 2% 5.89 6.41 mg/kg 9 Barium 165 165 100% 19.90 1160 127 30,000 5% 76.13 82.40 mg/kg 0 Beryllium 43 43 100% 0.15 0.97 1.2 2.3 0 응 0.51 0.56 mg/kg Cadmium 21.2 1.7 210 8 5% 0.87 165 63 888 0.08 0.97 mg/kg 6 Calcium 43 43 100% 2690 115000 62131 NA 14% 30832.37 45238.59 mg/kg Chromium 165 165 100% 2.30 676 37.8 2,000 24 15% 28.23 32.44 mg/kg Cobalt 43 43 100% 1.80 14.2 16.9 2,100 0 0 % 6.86 7.71 mg/kg 165 160 97% 4.40 1510 30.6 16,000 22 13% 23.06 25.97 Copper mg/kg 14415.35 15659.35 Iron 165 165 100% 2540 64700 34014 ID 4 2% mg/kg Lead 165 165 100% 6.40 820 59.8 400 48 29% 59.50 68.83 mg/kg Magnesium 43 43 100% 1050 59700 15014 5 12% 8606.23 11150.11 1,000,000 mg/kg Manganese 43 43 100% 66.80 15000 429 2,000 16 37% 836.13 1230.59 mg/kg 0.11 165 46 28% 0.10 1.4 0.25 130 14 9% 0.12 Mercury mg/kg Nickel 43 43 100% 4.50 46.7 50.9 32,000 0 0% 19.50 22.06 mg/kg 43 0 43 100% 370.00 4830 6757 0 응 1836.00 2082.29 Potassium NA mg/kg Selenium 165 22 13% 0.41 3.8 1.64 2,100 2 1% 0.60 0.64 mg/kg 31 19 Silver 165 19% 0.07 2.3 0.14 2,000 12% 1.35 1.68 mg/kg

1,000,000

28

3,700

140,000

1

1

3

33

2%

2%

7%

20%

194.08

0.20

31.47

106.76

245.43

0.22

37.34 mg/kg

120.70 mg/kg

mg/kg

mg/kg

ID = Inadequate data to develop criterion; NA = Not available

65%

100%

100%

2%

113

0.44

9.50

9.90

936

189

5650

0.44

504

0.75

53.8

116

28

1

43

165

43

43

43

165

10/11/95

Sodium

Zinc

Thallium

Vanadium

TABLE 2

TABLE SUBSURFACE SOIL INORGANIC ANALYTICAL RESULTS LOWER ECORSE CREEK DUMP SITE WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | MDNR Residentia | l Site Specific | | | | |
|-----------|----------|------------|-----------|----------|----------|------------------|-----------------|----------------------|---------------------|---------|-------------------|
| | Total | Positive | Detection | Detected | Detected | Cleanup Criteria | a Background | Positive Detections | Detection Frequency | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Direct Contact | Concentrations | Exceeding Background | Exceeding Backgroun | d Mean | UCL 95% Units |
| Aluminum | 40 | 40 | 100% | 2760.00 | 19200 | ID | 25958 | 0 | 0% 1 | 0738.98 | 12362.20 mg/kg |
| Antimony | 36 | 30 | 83% | 0.00 | 4.0 | 150.0 | 0.45 | 11 | 31% | 0.59 | 0.87 mg/kg |
| Arsenic | 347 | 328 | 95% | 1.10 | 30.9 | 5.5 | 13.6 | 135 | 39% | 7.19 | 7.61 mg/kg |
| Barium | 347 | 347 | 100% | 10.80 | 603 | 30000 | 127 | 150 | 43% | 79.11 | 83.28 mg/kg |
| Beryllium | 40 | 39 | 98% | 0.18 | 1.3 | 2.3 | 1.2 | 1 | 3% | 0.59 | 0.66 mg/kg |
| Cadmium | 347 | 67 | 19% | 0.04 | 3.9 | 210.0 | 1.7 | 32 | 9% | 0.93 | $1.00~{ m mg/kg}$ |
| Calcium | 40 | 40 | 100% | 1250.00 | 64300 | NA | 62131 | 1 | 3% 2 | 5701.20 | 39030.16 mg/kg |
| Chromium | 347 | 335 | 97% | 4.00 | 499 | 2000 | 37.8 | 145 | 42% | 22.41 | 24.61 mg/kg |
| Cobalt | 40 | 40 | 100% | 2.40 | 18.1 | 2100.0 | 16.9 | 1 | 3% | 8.27 | 9.43 mg/kg |
| Copper | 347 | 326 | 94% | 5.00 | 810 | 16000 | 30.6 | 157 | 45% | 24.26 | 26.35 mg/kg |
| Iron | 347 | 347 | 100% | 2270.00 | 114000 | ID | 34014 | 143 | 41% 1 | 6980.52 | 17910.99 mg/kg |
| Lead | 347 | 346 | 100% | 0.64 | 4510 | 400 | 59.8 | 168 | 48% | 44.70 | 50.01 mg/kg |
| Magnesium | 40 | 40 | 100% | 648.00 | 18700 | 1000000 | 15014 | 3 | 8% | 7405.12 | 9649.19 mg/kg |
| Manganese | 40 | 40 | 100% | 65.80 | 2010 | 2000 | 429 | 3 | 8% | 274.05 | 341.79~mg/kg |
| Mercury | 347 | 87 | 25% | 0.10 | 8.8 | 130.0 | 0.25 | 60 | 17% | 0.15 | 0.16 mg/kg |
| Nickel | 40 | 40 | 100% | 5.00 | 156 | 32000 | 50.9 | 1 | 3% | 29.10 | 34.05 mg/kg |
| Potassium | 40 | 40 | 100% | 302.00 | 5100 | NA | 6757 | 0 | 0% | 2322.01 | 2801.01 mg/kg |
| Selenium | 347 | 34 | 10% | 0.56 | 6.1 | 2100.0 | 1.64 | 20 | 6% | 0.95 | $1.01~{ m mg/kg}$ |
| Silver | 347 | 17 | 5% | 0.08 | 12.7 | 2000.0 | 0.14 | 16 | 5% | 1.75 | 1.98 mg/kg |
| Sodium | 40 | 33 | 83% | 185.00 | 981 | 1000000 | 504 | 13 | 33% | 421.26 | 537.31 mg/kg |
| Thallium | 40 | 2 | 5% | 1.10 | 1.4 | 28.0 | 0.75 | 2 | 5% | 0.35 | 0.43~mg/kg |
| Vanadium | 40 | 40 | 100% | 7.60 | 55.6 | 3700.0 | 53.8 | 1 | 3% | 27.20 | |
| Zinc | 347 | 343 | 99% | 11.40 | 1090 | 140000 | 116 | 166 | 48% | 84.30 | 92.46 mg/kg |

ID = Inadequate data to develop criterion; NA = Not available
10/11/95

TABLE 3 SURFACE SOIL VOC ANALYTICAL RESULTS LOWER ECORSE CREEK DUMP SITE WYANDOTTE, MICHIGAN

Minimum Maximum MDNR Residential Total Positive Detection Detected Detected Cleanup Criteria Positive Detections Detection Frequency Detections Frequency Parameter Analyses Value Value Direct Contact Exceeding Criteria Exceeding Criteria Mean UCL 95% Units Chloromethane 46 0 0.00 0.0 200,000 ND 6.27 ua/ka 0왕 Bromomethane 46 0 0 응 0.00 0.0 150,000 0 0 % ND 6.27 ug/kg 0 0 % Vinyl Choride 46 0 응 0.00 0.0 1,200 ND 6.27 ug/kg 46 0 0% 0.0 670,000 0% 10.25 8.06 Chloroethane 0.00 ug/kg Methylene Chloride 46 12 26% 2.00 19.0 340,000 0 % 17.42 10.03 ug/kg Acetone 46 12 26% 5.00 76.0 11,000,000 Λ O& 4.5 6.41 ug/kg Carbon Disulfide 46 4 9% 2.00 10.0 12,000,000 0 % ND 6.27 ug/kg 1.1-Dichloroethane 0 46 0 응 0.00 0.0 110,000 0 % ND 6.27 ug/kg 1.1-Dichloroethane 46 0 N% 0.00 0.0 13,000,000 0 0 % ND 6.27 ug/kg 0 0% 1,2-Dichloroethane (total) 0.00 0.0 1200000.00 0 % ND 6.27 ug/kg Chloroform 46 0 N% 0.00 0.0 420,000 0% ND 6.27 ug/kg 1,2-Dichloroethane 46 Ω N% 0.00 0.0 28,000 O& ND 6.27 ug/kg 2 4% 2-Butanone 46 3.00 5.0 200,000,000 O& 4.0 6.28 ug/kg 2 1,1,1-Trichloroethane 46 4% 2.00 3.0 3,100,000 0 % 2.5 6.31 ug/kg Carbon Tetrachloride 0 46 0 응 0.00 0.0 20,000 0 % ND 6.27 ug/kg Bromodichloromethane 46 0 0% 0.00 0.0 41,000 0 0 % ND 6.27 ug/kg 0 1,2-Dichloropropane 46 0 응 0.00 0.0 38,000 0 % ND 6.27 ug/kg cls-1,3-Dichloropropene 0 N% 46 0.00 0.0 14,000 O& ND 6.27 ug/kg Trichloroethene 46 1 28 3.00 3.0 160,000 O& 3.0 6.27 ug/kg Dibromochloromethane 0 46 0 응 0.00 0.0 31,000 Λ 0 % ND 6.27 ug/kg 1,1,2-Trichloroethene 46 1 2% 3.00 3.0 45,000 0 % 3.0 6.27 ug/kg 0 Benzene 46 0 응 0.00 0.0 88,000 0 % ND 6.27 ug/kg Trans-1,3-Dichloropropene 0 0% 14,000 0 0 % 0.00 0.0 ND 6.27 ug/kg 0 Bromoform 46 N% 0.00 0.0 320,000 O& ND 6.27 ug/kg 0 N% 4-Methyl-2-Pentanone 46 0.00 0.0 5,500,000 O& ND 6.27 ug/kg 2-Hexanone 46 0 N% 0.00 0.0 15,000,000 0 % ND 6.27 ug/kg Tetrachloroethene 46 0 0 응 0.00 0.0 50,000 Ω 0 % ND 6.27 ug/kg 1,1,2,2-Tetrachloroethane 0 0% 0.00 0.0 13,000 0 % ND 6.27 ug/kg Toluene 46 28 2.00 2.0 24,000,000 O& 2.0 6.36 ug/kg 0 0 ug/kg Chlorobenzene 46 0 응 0.00 0.0 2,100,000 0 % ND 6.27 Ethylbenzene 46 0 N% 0.00 0.0 11,000,000 O& ND 6.27 ug/kg Styrene 46 0 N% 0.00 0.0 85,000 O& ND 6.27 ug/kg 0.00

200,000,000

O&

1.5

6.57

ug/kg

0.0

ND = Compound was Not Detected

46

Ω

N%

Xylenes (total)

TABLE 4
SURFACE SOIL VOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

Minimum Maximum MDNR Residential Total Positive Detection Detected Detected Cleanup Criteria Positive Detections Detection Frequency Mean UCL 95% Parameter Analyses Detections Frequency Value Value Direct Contact Exceeding Criteria Exceeding Criteria Units Chloromethane 145 0 0% 0.0 0.0 200,000 ND 12.67 ua/ka Bromomethane 145 0 0 응 0.0 0.0 150,000 0 0 % ND 12.67 ug/kg 0 0 % Vinyl Choride 145 0 응 0.0 0.0 1,200 ND 12.67 ug/kg 0 N% 11.0 0% 12.67 Chloroethane 145 110 670,000 ND ug/kg Methylene Chloride 145 29 20% 7.0 260 340,000 0 % 13.60 15.72 ug/kg Acetone 145 79 54% 1.0 1300 11,000,000 O& 92.01 129.43 ug/kg Carbon Disulfide 145 56 39% 1.0 65000 12,000,000 0 % 20.81 28.33 ug/kg 1.1-Dichloroethane 145 0 0.0 0 % 0.0 110,000 0 % ND 12.67 ug/kg 1.1-Dichloroethane 145 0 N% 0.0 13,000,000 0 0 % 0.0 ND 12.67 ug/kg 1,2-Dichloroethane (total) 145 0 응 0.0 0.0 120000.00 0 % ND 12.67 ug/kg Chloroform 145 1 1 % 2.0 2.0 420,000 O& 11.11 12.68 ug/kg 1,2-Dichloroethane 145 0 N% 0.0 330 28,000 0 % ND 12.67 ug/kg 54 37% 2-Butanone 145 2.0 190 200,000,000 0 % 21.45 26.25 ug/kg 1,1,1-Trichloroethane 145 0 0 % 0.0 0.0 3,100,000 0 % ND 12.67 ug/kg Carbon Tetrachloride 145 0.0 0 응 0.0 20,000 0 % ND 12.67 ug/kg Bromodichloromethane 145 0 0 응 0.0 0.0 41,000 0 0 % ND 12.67 ug/kg 0 1,2-Dichloropropane 145 0 % 0.0 0.0 38,000 0 % ND 12.67 ug/kg cls-1,3-Dichloropropene 0 O% 145 0.0 0.0 14,000 O& ND 12.67 ug/kg Trichloroethene 145 1 1% 3.0 0.0 160,000 O& 11.11 12.67 ug/kg Dibromochloromethane 0 145 0 응 0.0 0.0 31,000 Λ 0 % ND 12.67 ug/kg 1,1,2-Trichloroethene 145 0 0% 0.0 0.0 45,000 0 % ND 12.67 ug/kg Benzene 145 3% 0.7 26.0 88,000 0 % 10.70 12.19 ug/kg Trans-1,3-Dichloropropene 0 0% 0.0 14,000 0 0 % 145 0.0 ND 12.67 ug/kg Bromoform 145 N% 0.0 0.0 320,000 ND 12.67 ug/kg 0 4-Methyl-2-Pentanone 145 N% 0.0 0.0 5,500,000 O& ND 12.67 ug/kg 2-Hexanone 145 0 N% 0.0 0.0 15,000,000 0 % ND 12.67 ug/kg Tetrachloroethene 145 0 0 응 0.0 0.0 50,000 Ω 0 % ND 12.67 ug/kg 1,1,2,2-Tetrachloroethane 145 2 1% 2.0 13,000 0 % 12.75 10.0 11.10 ug/kg Toluene 145 4% 1.0 15.0 24,000,000 0 % 10.83 12.43 ug/kg 145 0 0 Chlorobenzene 0 응 0.0 2,100,000 0 % 12.67 ug/kg 13.0 ND Ethylbenzene 145 3 28 1.0 11.0 11,000,000 O& 10.80 12.30 ug/kg 1 2.0 Styrene 145 1% 94.0 85,000 0 % 11.10 12.67 ug/kg

4%

0.6

21.0

200,000,000

0 %

11.03

12.61 ug/kg

ND = Compound was Not Detected

145

6

Xylenes (total)

Table 5

SURFACE SOIL CYANIDE ANALYTICAL RESULTS FROM PREVIOUS INVESTIGATION LOWER ECORSE CREEK DUMP SITE WYANDOTTE, MICHIGAN

| Parameter | Total Analyses | Positive Detections | | Detected | Detected | Background | - | | Detection Frequency Exceeding Background | | on UCL 95% | Units |
|-----------|---|------------------------|-----------|----------|----------|----------------|------------------|----------------------|--|---------|---------------|-------|
| Cyanide | 120 | 88 | 73% | 0.55 | 1730 | 0.49 | 9,300 | 86 | 72% | 6.02 | 9.81 | mg/kg |
| | Table 6 SUBSURFACE SOIL CYANIDE ANALYTICAL RESULTS FROM PREVIOUS INVESTIGATION LOWER ECORSE CREEK DUMP SITE WYANDOTTE, MICHIGAN | | | | | | | | | | | |
| | | | | | | - | MDNR Residential | | | Mean | | |
| | Total | Positive | Detection | Detected | Detected | Background | Cleanup Criteria | Positive Detections | Detection Frequency E | Stimati | on | |
| Parameter | Analyses | Detections | Frequency | Value | Value (| Concentrations | s Direct Contact | Exceeding Background | Exceeding Background | (a) | UCL 95% | Units |
| Cyanide | 307 | 237 | 77% | 0.59 | 32300 | 0.49 | 9,300 | 237 | 77% | 593.70 | 1005.72 | mg/kg |

Table 7

SURFACE SOIL CYANIDE ANALYTICAL RESULTS FROM RI LOWER ECORSE CREEK DUMP SITE WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | Site Specific | MDNR Residential | | | Mean | | |
|-----------|---|------------|-----------|----------|----------|---------------|------------------|----------------------|-----------------------|--------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | l Background | Cleanup Criteria | Positive Detections | Detection Frequency E | stimat | ion | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Concentration | s Direct Contact | Exceeding Background | Exceeding Background | (a) | UCL 95% | Units |
| Cyanide | 218 | 39 | 18% | 0.24 | 4.00 | 0.49 | 9,300 | 10 | 5% | 0.18 | 0.19 | mg/kg |
| | Table 8 SURFACE SOIL CYANIDE ANALYTICAL RESULTS FROM RI LOWER ECORSE CREEK DUMP SITE WYANDOTTE, MICHIGAN | | | | | | | | | | | |
| | | | | Minimum | Maximum | Site Specific | MDNR Residential | | | Mean | | |
| | Total | Positive | Detection | Detected | Detected | l Background | Cleanup Criteria | Positive Detections | Detection Frequency E | stimat | ion | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Concentration | s Direct Contact | Exceeding Background | Exceeding Background | (a) | UCL 95% | Units |
| Cyanide | 59 | 5 | 9% | 0.37 | 1.50 | 0.49 | 9,300 | 2 | 3% | 0.23 | 0.26 | mg/kg |

< IMG SRC 0596304H>

< IMG SRC 0596304I>

TABLE 11
SURFACE SOIL DIOXIN ANALYTICAL RESULTS FROM RI
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | MDNR Residential Po | sitive Detection | s Detection Fre | auency | | |
|---------------------|----------|------------|-----------|----------|----------|---------------------|------------------|-----------------|---------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | Cleanup Criteria | Exceeding | Exceeding | 10.0007 | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Direct Contact | Criteria | Criteria | Mean | UCL 95% | Units |
| TCDFs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.51 | ng/g |
| 2,3,7,8-TCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.28 | ng/g |
| PeCDFs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.49 | ng/g |
| 1,2,3,7,8-PeCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.16 | ng/g |
| 2,3,4,7,8-PeCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.17 | ng/g |
| HxCDFs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.12 | ng/g |
| 1,2,3,4,7,8-HxCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.05 | ng/g |
| 1,2,3,6,7,8-HxCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.05 | ng/g |
| 2,3,4,6,7,8-HxCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.05 | ng/g |
| 1,2,3,7,8,9-HxCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.06 | ng/g |
| HpCDFs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.06 | ng/g |
| 1,2,3,4,6,7,8-HpCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.06 | ng/g |
| 1,2,3,4,7,8,9-HpCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.08 | ng/g |
| OCDF | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.07 | ng/g |
| RCDDs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.24 | ng/g |
| 2,3,7,8-TCDD | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.10 | ng/g |
| PeCDDs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.13 | ng/g |
| 1,2,3,7,8-PeCDD | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.13 | ng/g |
| HxCDDs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.09 | ng/g |
| 1,2,3,4,7,8-HxCDD | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.11 | ng/g |
| 1,2,3,6,7,8-HxCDD | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.09 | ng/g |
| 1,2,3,7,8,9-HxCDD | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.10 | ng/g |
| HpCDDs (total) | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.14 | ng/g |
| 1,2,3,4,6,7,8-HpCDD | 8 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.12 | ng/g |
| OCDD | 8 | 0 | 63% | 0.47 | 16.0 | | 0 | 0% | 7.35 | 1835.97 | ng/g |
| Total TCDDs equiv | | | | | 0.016 | 0.09 | 0 | 0% | | | |

TABLE 12
SURFACE SOIL DIOXIN ANALYTICAL RESULTS FROM RI
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | MDNR Residential Po | sitive Detections | s Detection Fre | quency | | |
|---------------------|----------|------------|-----------|----------|----------|---------------------|-------------------|-----------------|--------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | Cleanup Criteria | Exceeding | Exceeding | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Direct Contact | Criteria | Criteria | Mean | UCL 95% | Units |
| TCDFs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.22 | ng/g |
| 2,3,7,8-TCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.18 | ng/g |
| PeCDFs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.18 | ng/g |
| 1,2,3,7,8-PeCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.18 | ng/g |
| 2,3,4,7,8-PeCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.20 | ng/g |
| HxCDFs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.07 | ng/g |
| 1,2,3,4,7,8-HxCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.09 | ng/g |
| 1,2,3,6,7,8-HxCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.07 | ng/g |
| 2,3,4,6,7,8-HxCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.09 | ng/g |
| 1,2,3,7,8,9-HxCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.10 | ng/g |
| HpCDFs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.12 | ng/g |
| 1,2,3,4,6,7,8-HpCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.12 | ng/g |
| 1,2,3,4,7,8,9-HpCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.15 | ng/g |
| OCDF | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.12 | ng/g |
| RCDDs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.11 | ng/g |
| 2,3,7,8-TCDD | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.09 | ng/g |
| PeCDDs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.25 | ng/g |
| 1,2,3,7,8-PeCDD | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.25 | ng/g |
| HxCDDs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.18 | ng/g |
| 1,2,3,4,7,8-HxCDD | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.20 | ng/g |
| 1,2,3,6,7,8-HxCDD | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.18 | ng/g |
| 1,2,3,7,8,9-HxCDD | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.19 | ng/g |
| HpCDDs (total) | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.32 | ng/g |
| 1,2,3,4,6,7,8-HpCDD | 16 | 0 | 0% | 0.0 | 0.0 | | 0 | 0% | ND | 0.32 | ng/g |
| OCDD | 16 | 7 | 44% | 0.78 | 8.6 | | 0 | 0% | 3.62 | 38.65 | ng/g |
| Total TCDDs equiv | | | | | 0.0086 | 0.09 | 0 | 0% | | | |

TABLE 13
SURFACE SOIL PCB ANALYTICAL RESULTS FROM RI
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | MDNR Residential Po | sitive Detection | s Detection Fre | quency | | |
|--------------|----------|------------|-----------|----------|----------|---------------------|------------------|-----------------|--------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | Cleanup Criteria | Exceeding | Exceeding | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Direct Contact | Criteria | Criteria | Mean | UCL 95% | Units |
| Aroclor-1016 | 16 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 56.74 | ug/kg |
| Aroclor-1221 | 16 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 56.74 | ug/kg |
| Aroclor-1232 | 16 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 56.74 | ug/kg |
| Aroclor-1242 | 16 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 56.74 | ug/kg |
| Aroclor-1248 | 16 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 56.74 | ug/kg |
| Aroclor-1254 | 16 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 56.74 | ug/kg |
| Aroclor-1260 | 16 | 2 | 12% | 24.0 | 48.0 | 2,300 | 0 | 0% | 36.0 | 56.74 | ug/kg |

TABLE 14
SURFACE SOIL PCB ANALYTICAL RESULTS FROM RI
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | MDNR Residential Po | sitive Detection | s Detection Fre | quency | | |
|--------------|----------|------------|-----------|----------|----------|---------------------|------------------|-----------------|--------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | Cleanup Criteria | Exceeding | Exceeding | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Direct Contact | Criteria | Criteria | Mean | UCL 95% | Units |
| Aroclor-1016 | 17 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 37.37 | ug/kg |
| Aroclor-1221 | 17 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 74.51 | ug/kg |
| Aroclor-1232 | 17 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 37.37 | ug/kg |
| Aroclor-1242 | 17 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 37.37 | ug/kg |
| Aroclor-1248 | 17 | 0 | 0% | 0.0 | 0.0 | 2,300 | 0 | 0% | ND | 37.37 | ug/kg |
| Aroclor-1254 | 17 | 1 | 6% | 250.0 | 250 | 2,300 | 0 | 0% | 55.58 | 106.67 | ug/kg |
| Aroclor-1260 | 17 | 1 | 6% | 4.0 | 4.0 | 2,300 | 0 | 0% | 29.92 | 42.31 | ug/kg |

TABLE 15
PERCHED GROUNDWATER INORGANIC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | Site | | | |
|-----------|----------|------------|-----------|----------|----------|------------|-----------|--------------|-------|
| | Total | Positive | Detection | Detected | Detected | Specific | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Background | Mean | UCL 95% | Units |
| Cyanide | 7 | 4 | 57% | 1.51 | 31.4 | 490 | 15.11 | 79.89 | ug/l |
| Aluminum | 3 | 0 | 0% | 0.0 | 0.0 | 25,958,000 | ND | 126.53 | ug/l |
| Antimony | 3 | 0 | 0% | 0.0 | 0.0 | 450 | ND | NA | ug/l |
| Arsenic | 3 | 1 | 33% | 5.6 | 5.6 | 13,640 | 2.96 | 29241220.78 | ug/l |
| Barium | 3 | 3 | 100% | 253.0 | 276 | 127,000 | 264.75 | NA | ug/l |
| Beryllium | 3 | 0 | 0% | 0.0 | 0.0 | 1,200 | ND | NA | ug/l |
| Cadmium | 3 | 0 | 0% | 0.0 | 0.0 | 1,650 | ND | NA | ug/l |
| Calcium | 3 | 3 | 100% | 236000 | 254000 | 62,131,000 | 247401.47 | NA | ug/l |
| Chromium | 3 | 0 | 0% | 0.0 | 0.0 | 37,800 | ND | 0.58 | ug/l |
| Cobalt | 3 | 3 | 100% | 0.6 | 1.6 | 16,900 | 1.19 | 13.12 | ug/l |
| Copper | 3 | 2 | 67% | 0.4 | 0.9 | 30,600 | 0.55 | 812.54 | ug/l |
| Iron | 3 | 3 | 100% | 1190 | 7790 | 34,014,000 | 7036.97 | 120902672.79 | ug/l |
| Lead | 3 | 0 | 0% | 0.0 | 0.0 | 59,800 | ND | NA | ug/l |
| Magnesium | 3 | 3 | 100% | 104000 | 129000 | 15,014,000 | 120647.89 | 153006.19 | ug/l |
| Manganese | 3 | 3 | 100% | 993.0 | 1420 | 429,000 | 1284.00 | 2067.72 | ug/l |
| Mercury | 3 | 0 | 0% | 0.0 | 0.0 | 250 | ND | NA | ug/l |
| Nickel | 3 | 1 | 33% | 4.2 | 4.2 | 50,900 | 2.24 | 39837.62 | ug/l |
| Potassium | 3 | 3 | 100% | 7420 | 25600 | 6,757,000 | 14586.39 | 328723.07 | ug/l |
| Selenium | 3 | 2 | 67% | 4.6 | 8.1 | 1,640 | 5.45 | 1853.04 | ug/l |
| Silver | 3 | 0 | 0% | 0.0 | 0.0 | 140 | ND | NA | ug/l |
| Sodium | 3 | 3 | 100% | 84700 | 100000 | 504,000 | 94537.47 | 113017.73 | ug/l |
| Thallium | 1 | 1 | 100% | 11.9 | 11.9 | 750 | 11.90 | NA | ug/l |
| Vanadium | 3 | 3 | 100% | 0.58 | 2.5 | 53,800 | 1.39 | 455.45 | ug/l |
| Zinc | 3 | 3 | 100% | 0.54 | 2.4 | 116,000 | 1.72 | 507.16 | ug/l |

ND = Analyte was Not Detected



TABLE 16
PERCHED GROUNDWATER VOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MICHIGAN

| | | | | Minimum | Maximum | | | |
|----------------------------|----------|------------|-----------|----------|----------|------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Chloromethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Bromomethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Vinyl Chloride | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chloroethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Methylene Chloride | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Acetone | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Carbon Disulfide | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1-Dichloroethene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1-Dichloroethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,2-Dichloroethene (total) | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chloroform | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,2-Dichloroethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 2-Butanone | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1,1-Trichloroethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Carbon Tetrachloride | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Bromodichloromethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,2-Dichloropropane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| cis-1,3-Dichloropropene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Trichloroethene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Dibromochloromethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1,2-Trichloroethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Trans-1,3-Dichloropropene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Bromoform. | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 4-Methyl-2-Pentanone | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 2-Hexanone | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Tetrachloroethene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1,2,2-Tetrachloroethane | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Toluene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chlorobenzene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Ethylbenzene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Styrene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Xylenes (total) | 2 | 1 | 50% | 1.00 | 1.00 | 1.00 |) NA | ug/L |

Table 17
PERCHED GROUNDWATER SVOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | Minimum | Maximum | | | |
|---------------------------|----------|------------|-----------|----------|----------|------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Naphthalene | 2 | 1 | 50% | 1.00 | 1.00 | 1.00 | NA | ug/L |
| Acenaphthylene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Acenaphthene | 2 | 1 | 50% | 0.50 | 0.50 | 0.50 | NA | ug/L |
| Fluorene, | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Phenanthrene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Anthracene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Fluoranthene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Pyrene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(a)anthraceno | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chrysene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| bis(2-Ethylhexy)phthalato | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(b)fluoranthene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(k)fluoranthene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(a)pyrene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Indeno(1,2,3-cd)pyrene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Dibenzo(a,h)anthracene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 2-Methylnaphthalene | 2 | 1 | 50% | 0.30 | 0.00 | 0.30 | NA | ug/L |
| Benzo(g,h,i)perylene | 2 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |

Table 18
SURFACE WATER INORGANIC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | , | | | | |
|-----------|----------|------------|-----------|-----------|-----------|-----------|-----------|-------|
| | | | | Minimum | Maximum | | | |
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Cyanide | 6 | 6 | 100% | 18.00 | 22.80 | 20.20 | 21.71 | ug/L |
| Aluminum | 6 | 6 | 100% | 848.00 | 2600.00 | 1582.00 | 2966.80 | ug/L |
| Antimony | 6 | 0 | 0% | 0.00 | 0.00 | ND | 2.18 | ug/L |
| Arsenic | 6 | 5 | 83% | 1.70 | 3.00 | 2,26 | 2.78 | ug/L |
| Barium | 6 | 6 | 100% | 80.40 | 108.00 | 89.73 | 98.64 | ug/L |
| Beryllium | 6 | 1 | 17% | 0.11 | 0.11 | 0.11 | 0.13 | ug/L |
| Cadmium | 6 | 5 | 83% | 0.59 | 1.70 | 0.92 | 1.74 | ug/L |
| Calcium | 6 | 6 | 100% | 72400.00 | 84600.00 | 77283.33 | 81296.98 | ug/L |
| Chromium | 6 | 6 | 100% | 7.70 | 19.30 | 11.23 | 15.64 | ug/L |
| Cobalt | 6 | 5 | 83% | 0.52 | 1.60 | 1.12 | 2.20 | ug/L |
| Copper | 6 | 6 | 100% | 7.30 | 27.10 | 13.48 | 24.06 | ug/L |
| Iron | 6 | 6 | 100% | 1320.00 | 4920.00 | 2565.00 | 4833.73 | ug/L |
| Lead | 6 | 4 | 67% | 19.80 | 57.80 | 30.08 | 46.07 | ug/L |
| Magnesium | 6 | 6 | 100% | 12400.00 | 14900.00 | 13666.67 | 14497.32 | ug/L |
| Manganese | 6 | 6 | 100% | 129.00 | 221.00 | 161.17 | 193.94 | ug/L |
| Mercury | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Nickel | 6 | 1 | 17% | 8.20 | 8.20 | 8.20 | 7.27 | ug/L |
| Potassium | 6 | 6 | 100% | 13800.00 | 18300.00 | 15616.67 | 17042.40 | ug/L |
| Selenium | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Silver | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Sodium | 6 | 6 | 100% | 702000.00 | 962000.00 | 824833.33 | 917865.95 | ug/L |
| Thallium | 0 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Vanadium | 6 | 6 | 100% | 4.20 | 9.80 | 6.43 | 9.14 | ug/L |
| Zinc | 6 | 3 | 50% | 102.00 | 211.00 | 139.67 | 166.11 | ug/L |
| | | | | | | | | |



Table 19
SURFACE WATER VOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | Minimum | Maximum | | | |
|----------------------------|----------|------------|-----------|----------|----------|-------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Chloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Bromomethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Vinyl Chloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Methylene Chloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Acetone | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Carbon Disulfide | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1-Dichloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1-Dichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,2-Dichloroethene (total) | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chloroform | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,2-Dichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 2-Butanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1,1-Trichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Carbon Tetrachloride | 6 | 1 | 17% | 1.00 | 1.00 | 1.00 | 12.05 | ug/L |
| Bromodichloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,2-Dichloropropane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| cis-1,3-Dichloropropene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Trichloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Dibromochloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1,2-Trichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzene | 6 | 1 | 17% | 2.00 | 2.00 | 2.00 | 6.89 | ug/L |
| Trans-1,3-Dichloropropene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Bromoform | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 4-Methyl-2-Pentanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 2-Hexanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Tetrachloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| 1,1,2,2-Tetrachloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Toluene | 6 | 4 | 67% | 4.00 | 9.00 | ND | 7.84 | ug/L |
| Chlorobenzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Ethylbenzene | 6 | 1 | 17% | 2.00 | 2.00 | 2.00 | 6.89 | ug/L |
| Styrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Xylenes (total) | 6 | 4 | 67% | 4.00 | 15.00 | 10.25 | 17.89 | ug/L |

Table 20
SURFACE WATER SVOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | Minimum Maximum | | | | |
|---------------------------|----------|------------|-----------|-----------------|----------|------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Napthalene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Acenaphthylene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Acenaphthene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Fluorene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Phenanthrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Anthracene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Fluoranthene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Pyrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(a)anthracene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Chrysene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| bis(2-Ethylhexy)phthalate | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(b)fluoranthene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(k)fluoranthene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(a)pyrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Indeno(1,2,3-cd)pyrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Dibenzo(a,h)anthracene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |
| Benzo(g,h,i)perylene | 6 | 0 | 0% | 0.00 | 0.00 | ND | ND | ug/L |

Table 21
SEDIMENT INORGANIC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| Parameter | Total Analyses | Positive Detections | Detection Frequency | Minimum Detected Value | Maximum Detected Value | Site Specific Background Concentrations | Positive Detection Exceeding Background | Detection Fr Exceedin Backgroun | a | UCL 95% |
|--------------------------|-------------------|------------------------|------------------------|------------------------------|------------------------------|---|---|---------------------------------------|----------|----------|
| Units Cyanide | 7 | 4 | 57% | 0.32 | 0.66 | 0.49 | 1 | 14% | 0.48 | 0.64 |
| mg/kg | | | | | | | | | | |
| Aluminum mg/kg | 7 | 7 | 100% | 9030.00 | 15300.00 | 25958.00 | 0 | 0% | 12347.14 | 14345.67 |
| Antimony | 7 | 6 | 86% | 0.68 | 1.30 | 0.45 | 6 | 86% | 0.90 | 1.62 |
| mg/kg Arsenic | 7 | 7 | 100% | 8.90 | 12.60 | 13.64 | 0 | 0% | 10.39 | 11.36 |
| mg/kg Barium | 7 | 7 | 100% | 153.00 | 217.00 | 127.00 | 7 | 100% | 180.57 | 199.32 |
| mg/kg Beryllium | 7 | 7 | 100% | 0.62 | 0.93 | 1.20 | 0 | 0% | 0.79 | 0.88 |
| mg/kg | | | | | | | | | | |
| Cadmium mg/kg | 7 | 7 | 100% | 2.30 | 4.60 | 1.65 | 7 | 100% | 3.61 | 4.54 |
| Calcium | 7 | 7 | 100% | 34400.00 | 61300.00 | 62131.00 | 0 | 0% | 50442.86 | 58490.73 |
| mg/kg Chromium | 7 | 7 | 100% | 48.60 | 120.00 | 37.80 | 7 | 100% | 73.20 | 95.74 |
| mg/kg Cobalt | 7 | 7 | 100% | 6.60 | 11.60 | 16.90 | 0 | 0% | 9.19 | 10.67 |
| mg/kg | | | | | | | | | | |
| Copper mg/kg | 7 | 7 | 100% | 89.00 | 151.00 | 30.60 | 7 | 100% | 115.29 | 113.34 |
| Iron | 7 | 7 | 100% | 19300.00 | 32600.00 | 34014.00 | 0 | 0% | 26785.71 | 30856.47 |
| mg/kg Lead | 7 | 7 | 100% | 192.00 | 625.00 | 59.80 | 7 | 100% | 285.00 | 415.42 |
| mg/kg Magnesium | 7 | 7 | 100% | 10900.00 | 18000.00 | 15014.00 | 2 | 29% | 13928.57 | 16142.41 |
| mg/kg | , | , | 100% | 10900.00 | 18000.00 | 13014.00 | 2 | 29% | 13920.37 | 10142.41 |
| Manganese mg/kg | 7 | 7 | 100% | 285.00 | 676.00 | 429.00 | 6 | 86% | 556.43 | 728.62 |
| Mercury | 7 | 7 | 100% | 0.38 | 8.90 | 0.25 | 7 | 100% | 1.86 | 9.29 |
| mg/kg Nickel mg/kg | 7 | 7 | 100% | 26.90 | 36.60 | 50.90 | 0 | 0% | 30.93 | 33.22 |
| mg/rg | | | | | | | | | | |

| Potassium mg/kg | 7 | 7 | 100% | 1850.00 | 3470.00 | 6757.00 | 0 | 0% | 2611.43 | 3306.28 |
|--------------------|---|---|------|---------|---------|---------|---|------|---------|---------|
| Selenium mg/kg | 7 | 7 | 100% | 1.30 | 2.40 | 1.64 | 3 | 43% | 1.63 | 1.95 |
| Silver mg/kg | 7 | 7 | 100% | 1.00 | 2.90 | 0.14 | 7 | 100% | 1.63 | 2.37 |
| Sodium mg/kg | 7 | 7 | 100% | 817.00 | 2170.00 | 504.00 | 7 | 100% | 1239.29 | 1659.97 |
| Thallium mg/kg | 7 | 0 | 0% | 0.00 | 0.00 | 0.75 | 0 | 0% | ND | 0.51 |
| Vanadium mg/kg | 7 | 7 | 100% | 29.70 | 46.20 | 53.80 | 0 | 0% | 37.17 | 41.62 |
| Zinc mg/kg | 7 | 7 | 100% | 457.00 | 685.00 | 116.00 | 7 | 100% | 582.57 | 664.98 |

Table 22
SEDIMENT VOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | *************************************** | , | | | | | |
|----------------------------|----------|---|-----------|----------|----------|-------|---------|-------|
| | | | | Minimum | Maximum | | | |
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Chloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Bromomethane | 6 | 0 | 0% | 0.00 | 0.00 | 0.00 | ND | ug/kg |
| Vinyl Chloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Chloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Methylene Chloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Acetone | 6 | 4 | 67% | 34.00 | 92.00 | 52.75 | 114.63 | ug/kg |
| Carbon Disulfide | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,1-Dichloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,1-Dichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,2-Dichloroethene (total) | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Chloroform | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,2-Dichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 2-Butanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,1,1-Trichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Carbon Tetrachloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Bromodichloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,2-Dichloropropane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| cis-1,3-Dichloropropene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Trichloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Dibromochloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,1,2-Trichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Benzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Trans-1,3-Dichloropropene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Bromoform | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 4-Methyl-2-Pentanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 2-Hexanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Tetrachloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Toluene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| 1,1,2,2-Tetrachloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Chlorobenzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Ethylbenzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Styrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 15.02 | ug/kg |
| Xylenes (total) | 6 | 1 | 17% | 10.00 | 10.00 | 10.00 | 15.54 | ug/kg |

NL = MDNR Residential Cleanup Criteria for Direct Contact was Not Listed

Table 23
SEDIMENT SVOC/PNA ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | Minimum | Maximum | | | |
|-----------------------------------|----------|------------|-----------|----------|----------|---------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Naphthalene | 7 | 0 | 0% | 0.00 | 0.00 | ND | 553.04 | ug/kg |
| Acenaphthylene | 7 | 1 | 14% | 91.00 | 91.00 | 91.00 | 802.26 | ug/kg |
| Acenaphthene | 7 | 0 | 0% | 0.00 | 0.00 | ND | 553.04 | ug/kg |
| Fluorene | 7 | 4 | 57% | 100.00 | 150.00 | 130.00 | 493.14 | ug/kg |
| Phenanthrene | 6 | 6 | 100% | 520.00 | 1300.00 | 928.33 | 1292.72 | ug/kg |
| Anthracene | 6 | 6 | 100% | 110.00 | 240.00 | 175.00 | 228.28 | ug/kg |
| Carbazole | 6 | 6 | 100% | 100.00 | 200.00 | 140.00 | 175.63 | ug/kg |
| Fluoranthene | 7 | 7 | 100% | 1300.00 | 2900.00 | 1985.71 | 2561.27 | ug/kg |
| Pyrene | 7 | 7 | 100% | 1000.00 | 3200.00 | 2171.43 | 3504.03 | ug/kg |
| Butylbenzylphthalate | 6 | 6 | 100% | 130.00 | 280.00 | 210.00 | 293.70 | ug/kg |
| Benzo(a)anthracene | 7 | 7 | 100% | 370.00 | 1200.00 | 762.86 | 1093.90 | ug/kg |
| Chrysene | 6 | 6 | 100% | 710.00 | 1500.00 | 1095.00 | 1438.63 | ug/kg |
| bis(2-Ethylhexy)phthalate | 6 | 6 | 100% | 2000.00 | 4000.00 | 2866.67 | 3798.67 | ug/kg |
| Di-n-octylphthalate | 6 | 6 | 100% | 200.00 | 560.00 | 360.00 | 594.06 | ug/kg |
| Benzo(b)fluoranthene | 6 | 6 | 100% | 490.00 | 1800.00 | 1273.33 | 2341.84 | ug/kg |
| Benzo(k)fluoranthene | 6 | 6 | 100% | 610.00 | 2000.00 | 1195.00 | 1865.51 | ug/kg |
| Benzo(a)pyrene | 7 | 7 | 100% | 490.00 | 920.00 | 727.14 | 811.40 | ug/kg |
| <pre>lndeno(1,2,3-cd)pyrene</pre> | 7 | 7 | 100% | 200.00 | 640.00 | 415.71 | 675.37 | ug/kg |
| Dibenzo(a,h)anthracene | 7 | 7 | 100% | 48.00 | 260.00 | 165.43 | 329.43 | ug/kg |
| Benzo(g,h.i)perylene | 7 | 7 | 100% | 110.00 | 300.00 | 201.43 | 285.63 | ug/kg |
| | | | | | | | | |

NL = MDNR Residential Cleanup Criteria for Direct Contact was Not Listed

Table 24
BASEMENT SUMP WATER INORGANIC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | | Minimum | Maximum | | | |
|-----------|---|----------|------------|-----------|----------|----------|-----------|------------|---------------|
| | | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Cyanide | | 8 | 3 | 38% | 8.10 | 5150 | 313.30 | 416141.05 | ug/L |
| Aluminum | | 8 | 6 | 75% | 8.10 | 6170 | 5392.06 | 809795.57 | ug/L |
| Antimony | | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2.64 | ug/L |
| Arsenic | | 8 | 7 | 88% | 1.60 | 1960 | 6.01 | 127.88 | ug/L |
| Barium | | 8 | 7 | 88% | 15.70 | 63.40 | 44.61 | 70.71 | ug/L |
| Beryllium | | 8 | 0 | 0% | 0.00 | 0.00 | ND | 0.08 | ug/L |
| Cadmium | | 8 | 2 | 25% | 0.71 | 11.30 | 1.04 | 16.61 | ug/L |
| Calcium | | 8 | 8 | 100% | 36700.00 | 231000 | 107850.11 | 224673.46 | ug/L |
| Chromium | | 8 | 2 | 25% | 9.50 | 36.1 | 6.53 | 32.00 | ug/L |
| Cobalt | | 8 | 7 | 88% | 0.56 | 11.6 | 1.95 | 8.63 | ug/L |
| Copper | | 8 | 8 | 100% | 4.40 | 5740 | 520.37 | 114440.23 | ug/L |
| Iron | | 8 | 7 | 88% | 1060.00 | 67500 | 8009.72 | 135848.25 | ug/L |
| Lead | | 8 | 8 | 100% | 7.90 | 125 | 17.35 | 80.39 | ug/L |
| Magnesium | | 8 | 7 | 88% | 5060.00 | 56200 | 32238.24 | 118165.80 | ug/L |
| Manganese | | 8 | 7 | 88% | 33.40 | 632 | 241.22 | 2134.05 | ug/L |
| Mercury | | 8 | 1 | 13% | 141.00 | 141 | 6.80 | 9689.38 | ug/L |
| Nickel | | 8 | 1 | 13% | 906.00 | 906 | 48.54 | 6765.26 | ug/L |
| Potassium | | 8 | 7 | 88% | 2480.00 | 9600 | 11064.53 | 410423.43 | ug/L |
| Selenium | | 8 | 2 | 25% | 4.40 | 14.1 | 3.37 | 8.00 | ug/L |
| Silver | | 8 | 2 | 25% | 0.54 | 2.40 | 0.51 | 1.24 | ug/L |
| Sodium | | 8 | 7 | 88% | 11700.00 | 2300000 | 198463.12 | 4244492.12 | ug/L Vanadium |
| | 8 | 8 | 100% | 0.98 | 16.4 | 5.01 | 15.24 | ug/L | |
| Zinc | | 8 | 1 | 13% | 26500.00 | 26500 | 1296.38 | 1783618.81 | ug/L |

Table 25
BASEMENT SUMP WATER VOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | **** | 1120112, 111 | | | | | |
|----------------------------|----------|------------|--------------|----------|----------|------|---------|-------|
| | | | | Minimum | Maximum | | | |
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Chloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Bromomethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Vinyl Chloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Chloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Methylene Chloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Acetone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Carbon Disulfide | 6 | 1 | 17% | 0.60 | 0.60 | 0.60 | 13.51 | ug/L |
| 1,1-Dichloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,1-Dichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,2-Dichloroethene (total) | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Chloroform | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,2-Dichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 2-Butanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,1,1-Trichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Carbon Tetrachloride | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Bromodichloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,2-Dichloropropane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| cis-1,3-Dichloropropene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Trichloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Dibromochloromethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,1,2-Trichloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Benzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Trans-1,3-Dichloropropene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Bromoform | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 4-Methyl-2-Pentanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 2-Hexanone | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Tetrachloroethene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| 1,1,2,2-Tetrachloroethane | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Toluene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Chlorobenzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Ethylbenzene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Styrene | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| Xylenes (total) | 6 | 0 | 0% | 0.00 | 0.00 | ND | 12.29 | ug/L |
| | | | | | | | | |

Table 26
BASEMENT SUMP WATER SVOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | , | | | | | |
|-----------------------------------|----------|------------|-----------|----------|----------|--------|---------|-------|
| | | | | Minimum | Maximum | | | |
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Napthalene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 33.31 | ug/L |
| Acenaphthylene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Acenaphthene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Fluorene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Phenanthrene | 8 | 1 | 13% | 17.00 | 17.0 | 13.99 | 19.10 | ug/L |
| Fluoranthene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Pyrene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Benzo(a)anthracene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| bis(2-Ethylhexy)phthalate | 8 | 1 | 13% | 140.00 | 140 | 121.72 | 144.96 | ug/L |
| Benzo(b)fluoranthene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Benzo(k)fluoranthene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Benzo(a)pyrene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| <pre>lndeno(1,2,3-cd)pyrene</pre> | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Dibenzo(a,h)anthracene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| Benzo(g.h,i)perylene | 8 | 0 | 0% | 0.00 | 0.0 | ND | 2202 | ug/L |
| 4-Methylphenol | 8 | 2 | 25% | 13.00 | 13.00 | 13.00 | N/A | ug/L |
| Phenol | 8 | 1 | 13% | 110.00 | 110.0 | 105.12 | 110 | ug/L |
| 2,4-Dimethylphenol | 8 | 1 | 13% | 26.00 | 26 | 20.26 | 40.7 | ug/L |
| Di-n-butylphthalate | 8 | 1 | 13% | 3.00 | 3.0 | 374.61 | 448504 | ug/L |
| Butylbenzylphthalate | 8 | 2 | 25% | 12.00 | 15.0 | 13.58 | 15.25 | ug/L |
| Di-n-octylphthalate | 8 | 1 | 13% | 78.00 | 78 | 89.69 | 102.04 | ug/L |
| Diethylphthalate | 8 | 1 | 13% | 12.00 | 12.00 | 106.58 | 1565 | ug/L |
| | | | | | | | | |

Table 27
BASEMENT SEDIMENT INORGANIC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | | | Minimum | Maximum | Site Specific 1 | Positive Detections | Detection Fre | equency | |
|--------------------|----------|------------|-----------|----------|----------|-----------------|---------------------|---------------|----------|----------------|
| | Total | Positive | Detection | | Detected | Background | Exceeding | Exceeding | , | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Concentrations | Background | Backgrour | nd Mean | UCL 95% |
| Units | | • | F.0.0 | 0 61 | 0.74 | 0.40 | • | 500 | 0.45 | 2 50 |
| Cyanide | 4 | 2 | 50% | 0.61 | 0.74 | 0.49 | 2 | 50% | 0.47 | 3.52 |
| mg/kg Aluminum | 4 | 4 | 100% | 8520.00 | 12500 | 25958.00 | 0 | 0% | 10761.69 | 13944 |
| mg/kg | 4 | 4 | 100% | 6520.00 | 12500 | 25956.00 | O | 0% | 10/61.69 | 13944 |
| Antimony | 3 | 2 | 67% | 0.68 | 6.50 | 0.45 | 2 | 67% | 5 09 | 65826296990130 |
| mg/kg | J | 2 | 0,70 | 0.00 | 0.50 | 0.13 | - | 0,70 | 3.05 | 03020290990130 |
| Arsenic | 4 | 4 | 100% | 5.50 | 6.70 | 13.64 | 0 | 0% | 6.06 | 6.87 |
| mg/kg | | | | | | | | | | |
| Barium | 4 | 4 | 100% | 84.20 | 117.00 | 127.00 | 1 | 25% | 115.57 | 202.54 |
| mg/kg | | | | | | | | | | |
| Beryllium | 4 | 4 | 100% | 0.26 | 0.61 | 1.20 | 0 | 0% | 0.44 | 0.87 |
| mg/kg | _ | _ | | | | | _ | | | |
| Cadmium | 4 | 4 | 100% | 0.45 | 5.80 | 1.65 | 2 | 50% | 3.05 | 186.90 |
| mg/kg Calcium | 4 | 4 | 100% | 25200.00 | 99200 | 62131.00 | 3 | 75% | 76323.53 | 345200.64 |
| mg/kg | 4 | 4 | 100% | 25200.00 | 99200 | 02131.00 | 3 | /56 | 70323.53 | 345200.04 |
| Chromium | 4 | 4 | 100% | 16.10 | 68.10 | 37.80 | 1 | 25% | 33.75 | 153.46 |
| mg/kg | - | - | 1000 | 10.10 | 00.10 | 37.00 | - | 250 | 33.73 | 133.10 |
| Cobalt | 4 | 4 | 100% | 5.80 | 9.50 | 16.90 | 0 | 0% | 7.01 | 9.67 |
| mg/kg | | | | | | | | | | |
| Copper | 4 | 4 | 100% | 21.70 | 1180 | 30.60 | 3 | 75% | 686.48 | 12011630.49 |
| mg/kg | | | | | | | | | | |
| Iron | 4 | 4 | 100% | 18800.00 | 32000 | 34014.00 | 0 | 0% | 24588.85 | 33562.86 |
| mg/kg | | | | | | | | | | |
| Lead | 4 | 4 | 100% | 22.00 | 1570 | 59.80 | 3 | 75% | 923.84 | 19864780.73 |
| mg/kg Magnesium | 4 | 4 | 100% | 3340.00 | 7590 | 15014.00 | 0 | 0% | 5929.23 | 11423.57 |
| mg/kg | 4 | 4 | 100% | 3340.00 | 7590 | 13014.00 | U | 0% | 3929.23 | 11423.57 |
| Manganese | 4 | 4 | 100% | 1200.00 | 1420 | 429.00 | 4 | 100% | 1291.03 | 1431.54 |
| mg/kg | - | - | 1000 | 1200.00 | | 127.00 | - | 1000 | 1271.00 | 1101.01 |
| Mercury | 4 | 3 | 75% | 0.17 | 1.10 | 0.25 | 1 | 25% | 0.43 | 33.40 |
| mg/kg | | | | | | | | | | |
| Nickel | 4 | 4 | 100% | 16.10 | 243 | 50.90 | 1 | 25% | 83.54 | 10018.33 |
| mg/kg | | | | | | | | | | |
| | | | | | | | | | | |

| Potassium mg/kg | 4 | 4 | 100% | 1030.00 | 2230 | 6757.00 | 0 | 0% | 1728.31 | 2925.68 |
|--------------------|---|---|------|---------|--------|---------|---|------|---------|-----------|
| Selenium mg/kg | 4 | 1 | 25% | 1.10 | 1.10 | 164 | 0 | 0% | 0.52 | 3.22 |
| Silver mg/kg | 4 | 4 | 100% | 0.32 | 1.10 | 0.14 | 4 | 100% | 0.63 | 2.61 |
| Sodium mg/kg | 4 | 4 | 100% | 245.00 | 859.00 | 504.00 | 2 | 50% | 544.56 | 1620.68 |
| Thallium mg/kg | 4 | 0 | 0% | 0.00 | 0.00 | 0.75 | 0 | 0% | ND | 0.32 |
| Vanadium mg/kg | 4 | 4 | 100% | 11.50 | 25.10 | 53.80 | 0 | 0% | 21.04 | 40.90 |
| Zinc mg/kg | 4 | 4 | 100% | 79.20 | 1210 | 116.00 | 3 | 75% | 910.41 | 190279.68 |

Table 28
BASEMENT SEDIMENT VOC ANALYTICAL RESULTS
LOWER ECORSE CREEK DUMP SITE
WYANDOTTE, MI

| | | "" | , , | | | | | |
|----------------------------|----------|------------|-----------|----------|----------|-------|---------|-------|
| | | | | Minimum | Maximum | | | |
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Chloromethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Bromomethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Vinyl Chloride | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Chloroethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Methylene Chloride | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Acetone | 1 | 1 | 100% | 14.00 | 14.00 | 14.00 | N/A | ug/kg |
| Carbon Disulfide | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,1-Dichloroethene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,1-Dichloroethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,1-Dichloroethene (total) | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Chloroform | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,2-Dichloroethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 2-ButaNDne | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,1,1-Trichloroethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Carbon Tetrachloride | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Bromodichloromethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,2-Dichloropropane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| cis-1,3-Dichloropropene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Trichloroethene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Dibromochloromethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,1,2-Trichloroethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Benzene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Trans-1,3-Dichloropropene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Bromoform | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 4-Methyl-2-Pentanone | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 2-Hexanone | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Tetrachloroethene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Toluene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| 1,1,2,2-Tetrachloroethane | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Chlorobenzene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Ethylbenzene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Styrene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Xylenes (total) | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| | | | | | | | | |

NL = MDNR Residential Cleanup Criteria for Direct Contact was Not Liste

Table 29

TABLE 4-30

BASEMENT SEDIMENT SVOC/PNA ANALYTICAL RESULTS

LOWER ECORSE CREEK DUMP SITE

WYANDOTTE, MI

| | | | | Minimum | Maximum | | | |
|-----------------------------------|----------|------------|-----------|----------|----------|------|---------|-------|
| | Total | Positive | Detection | Detected | Detected | | | |
| Parameter | Analyses | Detections | Frequency | Value | Value | Mean | UCL 95% | Units |
| Napthalene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Acenaphthylene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Acenaphthene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Fluorene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Phenanthrene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Anthracene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Carbazole | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Fluoranthene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Pyrene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Butylbenzylphthalate | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Benzo(a)anthracene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Chrysene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| bis(2-Ethylhexy)phthalate | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Di-n-octylphthalate | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Benzo(b)fluoranthene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Benzo(k)fluoranthene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Benzo(a)pyrene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| <pre>lndeno(1,2,3-cd)pyrene</pre> | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Dibenzo(a,h)anthracene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |
| Benzo(g,h,i)perylene | 1 | 0 | 0% | 0.00 | 0.00 | ND | N/A | ug/kg |

NL = MDNR Residential Cleanup Criteria for Direct Contact was Not Listed

Table 30
Summary of Risks Associated with Exposures and Major Contributors
Lower Ecorse Creek Dump Site
Wyandotte, Michigan

| | | RME, | ADULT | CTE A | DULT | RME C | HILD | CTE | CHILD | | | | | | | | | | | | | | | | | | | | |
|---------------------------|-----------------|--------------|--------------|------------|------------|--------------|--------------------|------------|--------|--------------|-------|-----------|--|-----|-----|-----------|----------|------------|-------------|-------------|-------------|--------------|---------------|---------------------------------------|-------------------|---------------------------------------|-------------------|--------------|-------------|
| | EXPOSURE | CANCER | HAZARD | CANCER | HAZARD | CANCER | HAZARD | CANCER | HAZARD | MAJOR | | | | | | | | | | | | | | | | | | | |
| AREA | ROUTE | RISK | INDEX | RISK | INDEX | RISK | INDEX | RISK | INDEX | CONTRIBUTOR | | | | | | | | | | | | | | | | | | | |
| RESIDENTIAL | Ingestion | 1E-6 | 2E-3 | 7E-8 | 3E-4 | 3E-6 | 2E-2 | 9E-7 | 6E-3 | PNAs | | | | | | | | | | | | | | | | | | | |
| SUREACE | Inhalation | 1E-11 | 2E-6 | (a) | (a) | 1E-11 | 1E-5 | (a) | (a) | | | | | | | | | | | | | | | | | | | | |
| SOILS | Dermal | 5E-6 | 3E-2 | 1E-7 | 2E-3 | 2E-6 | 5E-2 | 2E-7 | 4E-3 | PCBs, Mn | | | | | | | | | | | | | | | | | | | |
| | TOTAL | 6E-6 | 3E-2 | 2E-7 | 3E-3 | 5E-6 | 7E-2 | 1E-6 | 9E-3 | | | | | | | | | | | | | | | | | | | | |
| RESIDENTIAL | Ingestion | 5E-6 | 4E-3 | 3E-7 | 7E-4 | 1E-5 | 4E-2 | 4E-6 | 1E-2 | PNAs, Sb | | | | | | | | | | | | | | | | | | | |
| SUBSURFACE | Inhalation | | 2E-3 | (a) | (a) | 2E-8 | 8E-3 | (a) | (a) | CS2 | | | | | | | | | | | | | | | | | | | |
| SOILS | Dermal | 3E-7 | 4E-3 | 4E-8 | 2E-3 | 2E-8 | 1E-3 | 8E-9 | 5E-4 | | | | | | | | | | | | | | | | | | | | |
| | TOTAL | 5E-6 | 1E-2 | 3E-7 | 2E-3 | 1E-5 | 5E-2 | 4E-6 | 1E-2 | | | | | | | | | | | | | | | | | | | | |
| HOT-SPOT | Ingestion | | 2E-2 | (a) | (a) | | 1E-1 | (a) | (a) | CN- | | | | | | | | | | | | | | | | | | | |
| SURFACE | Dermal | | 9E-3 | (a) | (a) | | 1E-2 | (a) | (a) | | | | | | | | | | | | | | | | | | | | |
| SOILS | TOTAL | | 2E-2 | (a) | (a) | | 2E-1 | (a) | (a) | | | | | | | | | | | | | | | | | | | | |
| HOT-SPOT | Ingestion | | 2E-1 | (a) | (a) | | 2E+0 | | 7E-1 | CN- | | | | | | | | | | | | | | | | | | | |
| SUBSURFACE | Dermal | | 1E-1 | (a) | (a) | | 2E-1 | | 2E-2 | | | | | | | | | | | | | | | | | | | | |
| SOILS | TOTAL | | 4E-1 | (a) | (a) | | 2E+0 | | 7E-1 | | | | | | | | | | | | | | | | | | | | |
| PARK | Ingestion | 6E-6 | 4E-2 | 4E-7 | 7E-3 | 1E-5 | 3E-1 | 5E-6 | 1E-1 | As, PNAS | | | | | | | | | | | | | | | | | | | |
| SURFACE | Inhalation | | 2E-3 | (a) | (a) | 9E-9 | 8E-3 | (a) | (a) | CS2 | | | | | | | | | | | | | | | | | | | |
| SOILS | Dermal | 5E-6 | 3E-2 | 2E-7 | 3E-3 | 2E-5 | 6E-1 | 2E-7 | 5E-3 | As, PCBs | | | | | | | | | | | | | | | | | | | |
| G110 E3 GE | TOTAL | 1E-5 | 7E-2 | 6E-7 | 1E-2 | 3E-5 | 9E-1 | 5E-6 | 1E-1 | | | | | | | | | | | | | | | | | | | | |
| SURFACE | Ingestion | 7E-8 | 3E-3 | (a) | (a) | 8E-8 | 1E-2 | (a) | (a) | As, Mn | 4 | | | | | | | | | | | | | | | | | | |
| WATER | Dermal | 1E-8 | 6E-4 | (a) | (a) | 7E-9 | 1E-3 | (a) | (a) | As, Mn, CC14 | 4 | | | | | | | | | | | | | | | | | | |
| CEDIMENTO | TOTAL | 8E-8 | 3E-3 | (a) | (a) | 8E-8 | 1E-2 | (a) | (a) | DNIA | | | | | | | | | | | | | | | | | | | |
| SEDIMENTS | Ingestion | 1E-7 | 2E-5 2E-4 | (a) | (a) | 3E-7 | 2E-4 | (a) | (a) | PNAs, Hg | | | | | | | | | | | | | | | | | | | |
| | Dermal TOTAL | 5E-9 1E-7 | 2E-4 2E-4 | (a) | (a) | 2E-9 3E-7 | 2E-9 2E-4 | (a) (a) | (a) | PNAs | | | | | | | | | | | | | | | | | | | |
| SUMP WATER | Dermal | 1E-7 1E-8 | 2E-4 9E-6 | (a) | (a) | | ∠E-4 ALUATED | (a) | (a) | OT EVALUATED | NOT E | EVALUATED | | NC | мот | ਮਾਂ∩ਧਾ ਦਾ | MOT EXAL | NOT EXTATT | NOT EXTATII | MOT EXALITA | MOT EXALITA | MOT EXTAILIA | NOT EXTATITAT | אווו די | ארטיי בינא דוואיי | אווו די | ארטיי בינא דוואיי | MOT EXALITAT | NOT EVALUAT |
| SUMP WATER SUMP SEDIMENTS | | 1E-8 8E-9 | 9E-6 3E-9 | (a) (a) | (a) (a) | | ALUATED ALUATED | | | OT EVALUATED | | EVALUATED | | _ | _ | - | - | | | | | | | | | | | | NOT EVALUAT |
| SOME SENTIMENTS | TOTAL | он-9 2E-8 | 3E-3 | (a) (a) | (a) (a) | | ALUATED | | | OT EVALUATED | | EVALUATED | | _ | _ | _ | - | | | | | | | | | | | | NOT EVALUAT |
| | TOTAL | ZE-0 | 3E-3 | (a) | (a) | MOT EA | ATONIED | | IN | OI EVALUATED | NOI E | AMPONIED | | INC | MOI | MOI E | NOI EVAL | MOI EAMIO | MOI EAWDON | NOI EVALUA | NOI EVALUA | NOI EVALUA | NOI EVALUA | NOI EVALUAT | NOI EVALUAT | NOI EVALUAT | NOI EVALUAT | NOI EVALUAT | NOI EVALUAT |

⁽a) CIF exposure scenario was not evaluated because RME exposure scenario risks/hazard indices were below NCP range of 1x10-4 to 1x10-6 or below 1.

Table 31
COCs Detected in Soils
Lower Ecorse Creek Site FS
Page 1 of 2

| | Maximum Concentrati | ion | Target Method |
|--------------------------|---------------------|----------|-----------------|
| Chemical | Detected | PRG | Detection Limit |
| | (ug/kg) | (ug/kg) | (ug/kg) |
| Indeno(1,2,3-cd)pyrene | 19000 | 1.40E+04 | 330 |
| Benzo (a) pyrene | 6400 | 1.40E+03 | 330 |
| Dibenzo (a,b) anthracene | 3800 | 1.40E+03 | 330 |
| Benzo (g,h,i) perytene | 8000 | 1.50E+06 | 330 |
| Benzo (b&k) fluoranthene | 15000 | 1.40E+04 | 330 |
| 1,4-Dichlorobenzene | 14 | 1.10E+05 | 10 |
| 2-Methylphenol | 87 | 5.50E+06 | 10 |
| 2,4-Dimethylphenol | 170 | 2.10E+07 | 330 |
| Pentachlorophenol | 930 | 8.20E+04 | 3,400 |
| PCBS | | | |
| Arochlor-1254 | 250 | 2.30E+03 | 330 |
| Arochlor-1260 | 48 | 2.30E+03 | 330 |
| Dioxins | | | |
| OCDD | 16 | 9.00E-02 | 0.001 |
| Inorganics | | | |
| Aluminum | 1.92E+07 | ID | 700 |
| Antimony | 7.00E+03 | 1.50E+05 | 500 |
| Arsenic | 3.09E+04 | 1.37E+04 | 100 |
| Barium | 1.16E+03 | 3.00E+07 | 1000 |
| Beryllium | 1.30E+03 | 2.30E+03 | 200 |
| Cadmium | 2.12E+04 | 2.10E+05 | 2.10E+05 |
| Calcium | 1.15E+08 | NA | NL |
| Chromium | 6.76E+05 | 2.00E+06 | 200 |
| Cobalt | 1.81E+04 | 2.10E+06 | 500 |
| Copper | 1.51E+06 | 1.60E+07 | 1.60E+07 |
| Cyanide | 3.23E+07 | 1.00E+06 | 500 |
| Iron | 1.14E+08 | ID | 2000 |
| Lead | 4.51E+06 | 4.00E+05 | 1000 |
| Magnesium | 5.97E+06 | 1.00E+09 | 3000 |
| Manganese | 1.50E+07 | 2.00E+06 | 2000 |
| Mercury | 8.80E+00 | 1.30E+05 | 100 |
| Nickel | 1.56E+05 | 3.20E+07 | 1000 |
| Potassium | 5.10E+06 | NA | NL |
| Selenium | 6.10E+03 | 2.10E+06 | 500 |
| Silver | 1.27E+04 | 2.00E+06 | 500 |
| Sodium | 9.81E+05 | 1.00E+09 | NA |
| Thallium | 1.40E+03 | 2.80E+04 | 500 |
| Vanadium | 1.89E+05 | 3.70E+06 | 1000 |
| Zinc | 5.65E+06 | 1.40E+08 | 1000 |

Bold type indicate compounds exceeding the compound-specific MDEQ Direct Contact Value Cleanup Criteria, is based on MDEQ Interim Environmental Response Division Operational Memorandum #8. Revision 4: Generic Residential Cleanup Criteria or Calculated Background

Table 32
for COCs Detected in Soils
Lower Ecorse Creek Site FS
Page 1 of 2

| | Maximum Concentration | | Target Method |
|----------------------------|-----------------------|----------|-----------------|
| Chemical | Detected | PRG | Detection Limit |
| | (ug/kg) | (ug/kg) | (ug/kg) |
| VOCs | , 3, 3, | . 3. 3, | . 3. 3, |
| Methylene chloride | 260 | 3.40E+05 | 10 |
| Acetone | 1300 | 1.10E+07 | 100 |
| Carbon disulfide | 65000 | 1.20E+07 | 100 |
| 2-Butanone | 190 | 2.00E+09 | 100 |
| 1,1,1-Trichloroethane | 3 | 3.10E+06 | 10 |
| Trichloroethene | 3 | 1.60E+05 | 10 |
| 1,1,2-Trichloroethane | 6 | 4.50E+04 | 10 |
| Toluene | 15 | 2.40E+07 | 10 |
| Chloroform | 2 | 4.20E+05 | 10 |
| Benzene | 26 | 8.80E+04 | 10 |
| 1,1,2,2-Tetrachloroethane | 10 | 1.30E+04 | 10 |
| Ethylbenzene | 11 | 1.10E+07 | 10 |
| Styrene | 94 | 8.50E+04 | 10 |
| Xylene | 21 | 2.00E+08 | 30 |
| SVOCs | | | |
| Phenol | 2200 | 6.60E+07 | 330 |
| 4-Methylphenol | 87 | 2.10E+06 | 330 |
| N-Nitroso-Di-n-Propylamine | 320 | 3.70E+02 | 330 |
| Naphthalene | 14000 | 1.50E+07 | 330 |
| 2-Methylnapthalene | 9600 | ID | 330 |
| Dimethylphthalate | 450 | 1.00E+09 | 330 |
| Acenaphthylene | 2500 | 1.50E+06 | 330 |
| Acenaphthene | 2700 | 7.60E+07 | 330 |
| Dibenzofuran | 11000 | ID | 330 |
| Diethylphthalate | 590 | 3.20E+09 | 330 |
| Fluorene | 17000 | 5.10E+07 | 330 |
| N-Nitrosodiphenylamine (1) | 5000 | 5.20E+05 | 330 |
| Hexachlorobenzene | 27 | 6.20E+03 | 330 |
| Phenanthrene | 94000 | 1.50E+06 | 330 |
| Anthracene | 23000 | 4.20E+08 | 330 |
| Carbazole | 9000 | NL | NL |
| Di-n-butylphthalate | 1300 | 5.10E+07 | 330 |
| Fluoranthene | 150000 | 5.10E+07 | 330 |
| Pyrene | 120000 | 3.20E+07 | 330 |
| Butylbenzylphthalate | 2900 | 6.80E+07 | 330 |
| Benzo (a) anthracene | 60,000 | 1.40E+04 | 330 |
| Chrysene | 84000 | 1.40E+06 | 330 |
| bis(2-Ethylhexyl)phthalate | 77000 | 7.00E+05 | 330 |
| Di-n-octylphthalate | 21 | 7.60E+06 | 330 |
| Benzo (b) fluoranthene | 72000 | 1.40E+04 | 330 |
| Benzo (k) fluoranthene | 3900 | 1.40E+05 | 330 |
| | | | |

Table 33 Soil Cleanup Levels

| Chemical | Maximum Concentration Detected (ug/kg) | Cleanup Level (ug/kg) |
|------------------------|--|-----------------------|
| Indeno(1,2,3-cd)Pyrene | 19,000 | 14,000 |
| Benzo(a)pyrene | 6,400 | 1,400 |
| Dibenzo(a,h)anthracene | 3,800 | 1,400 |
| Benzo(b)fluoranthene | 72,000 | 14,000 |
| Arsenic | 30,900 | 13,600 |
| Cyanide | 3.23E+07 | 1.00E+06 |
| Lead | 4.51E+06 | 4.00E+05 |
| Benzo(a)anthracene | 60,000 | 14,000 |

12.0 RESPONSIVENESS SUMMARY

The public participation requirements of CERCLA sections 113 (k) (2) (I-v) and 117 of CERCLA have been met during the remedy selection process. Section 113(k)(2)(B)(iv) and 117(b) of CERCLA requires the EPA to respond "...to each of the significant comments, criticisms, and new data submitted in written or oral presentations" on a proposed plan for a remedial action. The Responsiveness Summary addresses concerns expressed by the public, potentially responsible parties (PRPs), and governmental bodies in written and oral comments received by EPA and the State regarding the proposed remedy for the Lower Ecorse Creek Site.

Background

U.S. EPA issued a fact sheet to the public in December, 1993, at the beginning of the Remedial Investigation. The Agency also hosted a public meeting on January 13, 1994, to provide background information on the Lower Ecorse Creek site, explain the Superfund process, and provide details of the upcoming investigation. The remedial investigation was completed in February 1996, and 'in February, 1996, U.S. EPA issued a second fact sheet to summarize the results of the investigation.

The RI/FS reports and the Proposed Plan for the Lower Ecorse Creek site were released to the public for review in April 1996. Information repositories have been established at the following location: Bacon Memorial Library, 45 Vinewood Avenue, Wyandotte, Michigan, 48192. The Administrative Record has been made available to the public at the U.S. EPA Docket Room in Region V and at the information repository.

A public meeting was held on May 9, 1996, to discuss the FS and the Proposed Plan. At this meeting, representatives from the U.S. EPA and the Michigan Department of Environmental Quality answered questions about the Site and the remedial alternatives under consideration. Formal oral comments on the Proposed Plan were documented by a court reporter. A verbatim transcript of this public meeting has been placed in the information repositories and Administrative Record. Written comments were also accepted at this meeting. The meeting was attended by approximately 50 persons, including local residents and PRPs.

The FS and Proposed Plan were available for public comment from April 3 0, 1996 through May 29, 1996. Comments received during the public comment period and the U.S. EPA's responses to those comments are included in the attached Responsiveness Summary, which is a part of this ROD. Advertisements announcing the availability of the Proposed Plan, start of the comment period were published in the News Herald and Detroit News newspapers on May 5, 1996. A correction was subsequently published in the Detroit Freepress and Detroit New on May 8 and 9, 1996, the New Herald on May 5, 1996, and in the Heritage newspaper on June 1, 1996, to correct the date of the public meeting.

During the comment period, EPA received approximately 5 written submittals of comments and 5 oral comments concerning the proposed plan.

Summary of Significant Comments

- 1: One commenter stated that they strongly recommend Alternative 3 since it is the most effective; it would remove ail contaminated soil and is the least costly. They also stated that Alternative 2 is not preferred because it would not remove deep contaminated soil, and the application of deed restrictions to the property would be "disasterous". They further requested that the additional sampling and excavation at the Oak Street location be scheduled for the month of August when the restaurant that uses the lot for customer parking is closed..
- Response 1: U.S. EPA acknowledges the comment. The Agency will work with the owner of the Oak Street property and the management of the restaurant that uses the parking lot to try to accommodate their schedule as best we can.
- Comment 2: One commenter stated that they definitely agree with the recommendation for Alternative 3 because it is the only way for the neighborhood to get rid of the stigma associated with the contamination. They additionally stated that the cleanup should

begin immediately. They did not understand why they would have to wait until next year, and why their property needs to be sampled again.

Response 2: U.S. EPA acknowledges the comment. We are constantly working to expedite cleanups at Superfund sites. One of the things done at this site to make it move through the process more quickly was to utilize the removal cleanup to gather data which could then be used in the Remedial phase of the project, avoiding having to resample areas to get the appropriate quality of data. We will work throughout this cleanup to streamline the process and complete the project as quickly and efficiently as we can.

However, there are things we are required to complete before we can actually start the excavation project. At this site at least two steps must be taken. First, we must determine who is going to perform the cleanup. Either U.S. EPA will do the work using Superfund Trust Fund money, or the potentially responsible parties will use their own resources to perform the work. U.S. EPA is committed to trying to get those responsible for creating contamination problems to perform the cleanup. If they fail to step up and commit to undertake the action U.S. EPA will be forced to use money from the Superfund Trust Fund to perform the work. In the long run, where there is not an immediate threat to public health or the environment, it is beneficial to everyone to avoid unnecessarily relying on the Trust Fund to pay for the cleanup and having the potentially responsible parties undertake the work. That makes those funds available for projects where there in no private party to step forward to do the work, At this site we intend on approaching the parties to ask them to perform the cleanup. We hope to complete that negotiation process within a couple of months after the ROD is issued. If they refuse, we are prepared to use the Fund monies.

The second step to complete is the design phase of the project. This phase primarily consists of assembling all of the plans for carrying out the cleanup, Including work plans to detail how the excavation will take place, health and safety plans for the workers, sampling plans to help determine when excavations are complete, restoration plans for the properties being affected, to name a few. Documents also need to be assembled to prepare to acquire bids and hire contractors to do the work. This design phase can take at least 6 months to complete and is necessary to ensure that the cleanup is performed properly and will meet specifications. We will take whatever steps we can to expedite this phase of the project, but it is not anticipated that it will be completed before the winter of this year. If that is the case we probably would not begin the excavation until spring because of the frozen condition of the soil which can make excavation extremely difficult. In most cases it would be best to wait for spring so the grounds thaw and working conditions are more acceptable.

The properties need to be resampled so we can more accurately estimate how much soil requires removal. In many instances we only have one sample from one location indicating a potential contamination problem. We need to take more samples around that location to confirm whether there is a contamination problem and to define the extent of that contamination before deciding if action is required.

Comment 3: One commenter stated that as long as there is an EPA presence in the area property values will be subjected to a lower than warranted value. They go on to say that if Alternative 3 were to be implemented and completed the "cloud of doubt" would be removed from the area and property values would again represent actual market values. The health and welfare of the people in the area, as well as the City of Wyandotte would be protected.

Response 3: The comment is acknowledged.

Comment 4: One commenter asked about Ecorse Creek pollution and cleanup in the summer.

They stated that the smell coming from the creek in the summer is pretty bad, and to possibly have the high spots in the creek dredged to allow fresh water to flow back up the creek might be beneficial.

They also asked why the contaminated properties are just not bought by the City and developed into a golf course or park, as has been done at other contaminated areas.

- Response 4: U.S. EPA has sampled the creek, both upstream and adjacent to the Site, and have found contaminants in the sediments and surface water. However, the same contaminants found adjacent to the Site were found upstream, indicating that the contamination is not from the Site, but is probably from some other source.

 Therefore, the creek will not be addressed as part of this cleanup effort.
 - U.S. EPA has made a strong commitment to employing cleanups which are permanent solution to contamination problems. By leaving contamination in place the cost of the remedy would rise because of the costs associated with constructing a soil cover over the contamination, maintaining the cover system, and implementing deed restrictions to ensure no one digs on the property in the future. In addition, leaving the contamination in the middle of a residential area might have a continuing detrimental effect on property values. We believe excavation and offsite disposal is the most cost effective solution to the problem.
- Comment 5: One commenter stated that the Army Corps of Engineers ("U.S. ACE")has
 Jurisdiction waterward of the Ordinary High Water Mark, and in any adjacent
 wetlands for that portion of the Ecorse River which is east of the Detroit and
 Toledo Railroad tracks. If the site clean-up may affect the course, capacity, or
 condition of the Ecorse River Downstream of where the work is occurring, they
 advise that U.S. EPA contact the U.S. ACE Detroit office prior to commencement
 of work for possible permit requirements.
- Response 5: U.S. EPA will comply with any substative, applicable, or relevant and appropriate permit requirements the U.S. ACE may have for any work occurring on-site, and will obtain all necessary permits for work occurring off-site. We win coordinate with the appropriate office of the U.S. ACE.
- Comment 6: One commenter stated that they agree with Alternative 3. They stated that they tried to sell their home and the offer they received came in \$33,000 below the estimated value. They would prefer to get rid of the problem and be in the position someday getting back to their normal lives.
- Response 6: The comment is acknowledged.
- Comment 7: One commenter asked how hazardous is the surface water in the creek? He expressed concern over children consuming the water. He went on to express concern over the contaminants entering the creek via surface water run-off from the site.
- Response 7: Surface water run-off of is a potential migration pathway for contaminants from the site. Through the selected cleanup, and off-site disposal of the contaminated soils, the threat of site-related contaminants entering the creek should be nuinimized. It should be noted that the creek had been sampled as part of the Remedial Investigation and it was found that sources other than this Site are probably contributing contamination to the creek. Risks to children exposed to the surface water and sediments from the creek were-evaluated as part of the risk assessment process. No unacceptable risks were found. U.S. EPA defines an unacceptable risk as an increased risk greater than 1 x 10-4 to 1 x 10-6 of an individual getting cancer

(or 1 in 10,000 to 1, in 1,000,000). For non-carcinogens it is defined as a Hazard Index greater than 1.0. For children at this site the carcinogenic risk for exposure to surface water and sediments was calculated to be 8 x 10-8 and 3 x 10-7, respectively. The non-carcinogenic risk was 0.001 and 0.0001 for surface water and sediments, respectively.

- Comment 8: One commenter asked whether we know exactly which properties will be excavated.
- Response 8: At this point we know that 470/80, 471, and the empty lot north of 470/480 will be excavated. The park area will also be excavated. On these properties we have sufficient data to justify the need to excavate. Other properties along North Drive and the property on Oak Street require additional sampling to determine whether excavation is warranted. We anticipated completing that sampling by the end of the fall of 1996.
- Comment 9: One commenter asked if work was going to continue on his property on Oak Street while we go through the remedy selection process.
- Response 9: The work in question, excavating contaminated soils from underneath the back porch at the residence on Oak Street, was post-poned until the homeowner tore down the porch as he agreed. The commenter has been informed that he should remain in touch with the On-Scene Coordinator to keep them appraised of his schedule so the cleanup can be coordinated.

ADMINISTRATIVE RECORD INDEX

< IMG SRC 0596304J>

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FOR

NORTH DRIVE SITE Wyandotte, MI

January 11, 1994

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NORTH DRIVE/LOWER ECORSE CREEK DUMP SITE

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